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## OMNIBUS ESSENTIAL FISH HABITAT AMENDMENT 2 DRAFT ENVIRONMENTAL IMPACT STATEMENT

Appendix E - Synopsis of Closed Area Technical Team analysis of juvenile groundfish habitats and groundfish spawning areas

Synopsis of juvenile groundfish habitat and spawning analysis

Intentionally blank.

Note - this appendix is adapted from a memorandum provided to the New England Fishery Maangement Council's Scientific and Statistical Committee on May 10, 2013.

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## Analytical approach

Between January and April 2013, the Closed Area Technical Team developed an analysis of data to assist in identifying areas that more restrictive measures could reduce impacts on juvenile groundfish habitat and groundfish spawning. Instead of focusing on physical characteristics of the environment that might be damaged by fishing and could be suitable habitat for groundfish, the CATT took an approach that focuses on aggregations of small juvenile groundfish and large fully-mature groundfish.

The CATT made a few key decisions about how to focus the analysis to meet the objectives. First, the CATT decided that the primary data source it would use to analyze juvenile and mature groundfish distribution would be from the various fishery-independent surveys, conducted by NMFS and coastal states. Figure 6 shows the geographic distribution of the surveys used for this analysis. Certain other surveys, such as RSA surveys or the Canadian survey were not readily available. The NMFS, MA DMF, and ME/NH surveys were the most useful for identifying hotspots or clusters of large catches. The IBS (Industry Based Survey) cod survey was also suitable, but the spatial domain of the survey was limited. The IBS goosefish and yellowtail flounder surveys were potentially suitable and were included in the analysis, but the sampling density was low and the analysis yielded few hotspots.

One important issue with survey data that was recognized by the CATT and addressed was the apparent overdispersion and high amount of zero catch observations in the survey catch per tow data. As such, it was unlikely that the data would be suitable for parametric analysis embedded in the Getis-Ords G* (henceforth simply called G*) statistic, particularly when interpreting the pvalue to distinguish clusters of significantly high catches. Although the $\mathrm{G}^{*}$ statistics is valid using data that is not normally distributed, Zhang et al (2008) published a proof that the G* statistics are not accurate for overdispersed data. It is furthermore common practice to either use non-parametric tests or transform survey data before analysis. A Box-Cox procedure was applied in R and Systat to potentially identify a transformation yielding distributions that were approximately normal. None were satisfactory, including a log (or any other) transformation of $\mathrm{N}+1$.

The CATT explored the issue by running several trials with untransformed and transformed data, but in the end followed the advice of Dr. Brian Kinlan to adjust the data in a two-step (Hurdle model like approach) procedure to down weight catches on tows that occur in strata having higher numbers of zero catch tows. The catch per tow was multiplied by the proportion of nonzero catches in a stratum during each year and survey, before applying a log transformation. This procedure yielded normally distributed data, adjusted for the proportion of zero tows in a stratum (i.e. catches in strata having higher proportions of no-catch tows were down weighted relative to strata where the catches were more consistently non-zero).

Size ranges that approximate age $0 / 1$ were chosen by the CATT for the juvenile groundfish hotspot analysis. A size threshold was selected that included all of age 0 fish and about $90 \%$ of age 1 fish from regenerated age length keys for 2002-2012 for the spring and fall NMFS trawl surveys (Table 5). Size ranges derived from the spring survey were applied to measured groundfish for all spring and summer surveys. Size ranges derived from the fall survey were applied to measured groundfish for all fall and winter surveys The CATTs rationale for choosing these size thresholds was to key in on the smallest juvenile groundfish caught by the lined survey trawls, which are more likely to be associated with bottom habitat that could be
adversely affected by fishing. The thresholds were always smaller than the L20 for that species maturity ogive, which had been re-estimated for 2002-2012 (Table 4).

In general, the L80 on the re-estimated maturity ogives were generally within 5 cm of the L50 and if used as a threshold for spawners would have favored identification of hotspots of small spawners. Instead, the CATT chose to focus the analysis on larger spawners which were thought to be more likely to have mature spawning behavior, higher fecundity, and better egg viability. Large spawners were identified using a threshold that larger fish made up about $20 \%$ of the total biomass in the 2002-2012 NMFS trawl surveys. Since growth at this size is typically slower than at younger ages, a single threshold was applied in all seasons for each species (see Table 8).

These transformed data were used to perform the $\mathrm{G}^{*}$ hotspot analyses, following the steps outlined in Table 9. For each survey, species, and size range (juveniles and large spawners) a spatial autocorrelation analysis was performed to identify distances that had significant positive correlations. When they existed (see examples in Figure 20 to Figure 28), the first statistically significant peak was used to set the G* Zone of Indifference, defining the neighborhood that was considered for identifying clusters. At other times, there was no first peak in autocorrelation and the maximum peak was used instead. Generally, if there was no statistically significant spatial autocorrelation, the $G^{*}$ procedure also failed to identify any clusters or hotspots. The zone of indifference setting for each $\mathrm{G}^{*}$ analysis performed is listed in Table 10.

Two important choices or assumptions were made in the hotspot analysis. One of these choices is the neighborhood of tows considered to be a potential hotspot. There are a variety of choices ranging from a fixed distance, inverse distance weighting, to a zone of indifference (with inverse distance weighting). The choice made by the CATT after considerable sensitivity analysis was a zone of indifference determined by a local maximum ("first peak") spatial autocorrelation. Unlike a fixed distance application, the zone of indifference was valid for all tows because no tows had no neighboring tows, a key violation of a fixed distance model which frequently gave warnings using the survey data. Only significant ( $\mathrm{p}<=0.05$ ) hotspots with above average catches were selected for further use as a hotspot (see Figure 10; Map 1). No standard p-value is available to determine significance, although p-values less than 0.05 were examined as a sensitivity analysis. For redfish, the hotspots tended to contract to a more centralized location in the Western Gulf of Maine with lower p-values.

Since the ultimate purpose of this analysis is to identify areas where a reduction in fishing would reduce impacts on juvenile groundfish habitat and groundfish spawning, for a variety of large mesh groundfish species, the CATT needed a way to summarize the hotspots across species and in shapes that were amenable to combinations into area options. The hotspots for all surveys were summarized in $100 \mathrm{km2}$ grids, compatible with SASI model outputs.

Juvenile groundfish hotspots for each stock were given an importance weight (Table 1), a simple arithmetic sum of four factors: Stock vulnerability, sub-population characteristics, residency characteristics, and substrate affinity. Stock vulnerability was chosen as a measure of how close the stock biomass is to the target biomass, i.e. $\mathrm{B}_{\mathrm{msy}} / \mathrm{B}$. Stocks at the target had a value of 1 , while overfished stocks had a value of 2 or more. Sub-population characteristics, residency characteristics, and substrate affinity were assigned a score from 1 to 3 based on published information and EFH source documents. More details are provided in a difference SSC document. Vulnerability or characteristics that were unknown (UNK) or could not be assigned were given a mean score as a proxy value in the final weighting sum.

Hotspots, i.e. clusters of significantly above average catches, of large mature groundfish were given similar importance weights using the same factors as applied for juvenile groundfish, but without the substrate affinity classification (Table 2), because the CATT decided that other factors (water temperature, moon phase, etc.) were more important to spawning of many groundfish species than was substrate affinity. Stocks were excluded from the seasonal hotspot summary gridding during seasons when the stock was not spawning (Table 2).

These weighted hotspot results were then summed by season over all species to guide the CATT to design potential juvenile groundfish area management options. The characteristics of these areas as well as those proposed by the Habitat PDT and Oversight Committee were analyzed for the number of juvenile and large spawner groundfish hotspots, Z-infinity scores from the SASI model, species diversity, potential displacement of net fishery revenue, etc. Hotspot grids and potential areas were compared (Figure 11 to Figure 13) with presence of observed developing, ripe, and running ripe groundfish to verify their location with respect to observations of spawning condition fish. Similarly the CATT intends to compare egg distribution from the ECOMON project with the results of the hotspot analysis as verification and to refine the timing of potential spawning closures.
 applied to the gridded hotspots for each species shaded in red. Grey shaded rows designate species that are not managed by catch shares.

| Stock (Red cells indicate selected stocks for Option 3) | Juvenile size threshold Age 0 and 1 length (90th percentile, cm) | Length at $20 \%$ female maturity (cm) (reestimated by CATT) | Vulnerability of species (Bmsy/B) ${ }^{1}$ | Sub-populations ${ }^{2}$ | Residency ${ }^{3}$ | Substrate ${ }^{4}$ | Final Weighting Sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GB Cod | 24 (Sp), 34 (Fa) | 36 | 14.11 | 2 | 1 | 3 | 20.11 |
| GOM Cod | 24 (Sp), 34 (Fa) | 36 | 5.53 | 3 | 1 | 3 | 12.53 |
| GB Yellowtail Flounder | 13 (Sp), 15 (Fa) | 25 | 9.39 | 1 | 2 | 1 | 13.39 |
| CC/GOM Yellowtail Flounder | 13 (Sp), 15 (Fa) | 25 | 4.21 | 1 | 2 | 1 | 8.21 |
| SNE/MA Yellowtail Flounder | 13 (Sp), 15 (Fa) | 25 | 0.77 | 1 | 2 | 1 | 4.77 |
| GOM Winter Flounder | 18 (Sp), 28 (Fa) | 27 | UNK | UNK | 2 | 1 | 10.04 |
| GB Winter Flounder | 18 (Sp), 28 (Fa) | 27 | 1.22 | 3 | 2 | 1 | 7.22 |
| SNE/MA Winter Flounder | 18 (Sp), 28 (Fa) | 27 | 6.17 | 3 | 2 | 1 | 12.17 |
| White Hake | 34 (Sp), 39 (Fa) | 25 | 1.21 | UNK | 2 | 1 | 6.04 |
| GOM Haddock | 24 (Sp), 34 (Fa) | 28 | 1.71 | 1 | 1 | 3 | 6.71 |
| GB Haddock | 24 (Sp), 34 (Fa) | 28 | 0.75 | 1 | 1 | 3 | 5.75 |
| Witch Flounder | 20 (Sp), 19 (Fa) | 28 | 2.45 | 3 | 2 | 1 | 8.45 |
| American Plaice | 12 (Sp), 18 (Fa) | 24 | 1.70 | UNK | 1 | 1 | 5.54 |
| Pollock | 23 (Sp), 32 (Fa) | 39 | 0.46 | 2 | 2 | 2 | 6.46 |
| Acadian Redfish | 14 (Sp), 13 (Fa) | 19 | 0.76 | 1 | 2 | 3 | 6.76 |
| Atlantic Halibut | see winter flounder | NA | 28.82 | UNK | 2 | 2 | 34.66 |
| Ocean Pout | 29 | $29^{6}$ | 12.05 | UNK | 1 | 2 | 16.88 |
| Northern (GOM-GB) Windowpane Flounder | see yellowtail flounder | 18 | 3.48 | UNK | 2 | 1 | 8.31 |
| Southern (SNE-MA) <br> Windowpane Flounder | see yellowtail flounder | 18 | 0.69 | UNK | 2 | 1 | 5.52 |
| Atlantic Wolffish | 47 | $47^{7}$ | 3.48 | UNK | UNK | 2 | 8.99 |
| Sum |  |  |  |  |  |  | 208.52 |
| Mean |  |  | 5.21 | 1.83 | 1.68 | 1.70 | 10.43 |

${ }^{1}$ Either SSBmsy/SSB or Bmsy/B used depending on what is reported in the assessment
${ }^{2}$ Derived from Table 81 in Framework 48 or from NEFSC biological data. 1=no subpopulations, 2=some evidence, 3=known subpopulations
${ }^{3}$ Based on information in literature. $1=$ less resident, more migratory; $2=$ more resident, less migratory
${ }^{4}$ Based on information in literature. 1=almost exclusively in mud or sand substrates, $2=0$ ccur in a variety of substrates including gravels, $3=$ strong affinity for coarse or hard substrates
${ }^{5}$ Sums include a mean value for unknowns
${ }^{6}$ From O'Brien et al. (1993)
${ }^{7}$ From Templeman (1986)

Synopsis of juvenile groundfish habitat and spawning analysis
 sum was applied by season to the gridded hotspots for each species shaded in red. Grey shaded rows designate species that are not managed by catch shares.

| Stock | Large spawner threshold (20\% of total biomass) | Length at $\mathbf{8 0 \%}$ female maturity (cm) (reestimated by CATT) | Vulnerability of species (Bmsy/B) ${ }^{1}$ | Subpopulations ${ }^{2}$ | Residency ${ }^{3}$ | Final weighting Sum ${ }^{4}$ | Spring multiplier | Summer multiplier | Fall multiplier | Winter multiplier |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GB Cod | 75 | 52 | 14.11 | 2 | 1 | 17.1 | 1 | 1 | 0 | 1 |
| GOM Cod | 75 | 52 | 5.53 | 3 | 1 | 9.5 | 1 | 1 | 0 | 1 |
| GB Yellowtail Flounder | 40 | 30 | 9.39 | 1 | 2 | 12.4 | 1 | 0 | 0 | 0 |
| CC/GOM Yellowtail Flounder | 40 | 30 | 4.21 | 1 | 2 | 7.2 | 1 | 0 | 0 | 0 |
| SNE/MA Yellowtail Flounder | 40 | 30 | 0.77 | 1 | 2 | 3.8 | 1 | 0 | 0 | 0 |
| GOM Winter Flounder | 45 | 31 | UNK | UNK | 2 | 9.0 | 1 | 0 | 0 | 1 |
| GB Winter Flounder | 45 | 31 | 1.22 | 3 | 2 | 6.2 | 1 | 0 | 0 | 1 |
| SNE/MA Winter Flounder | 45 | 31 | 6.17 | 3 | 2 | 11.2 | 1 | 0 | 0 | 1 |
| White Hake | 75 | 45 | 1.21 | UNK | 2 | 5.0 | 1 | 0 | 0 | 0 |
| GOM Haddock | 50 | 40 | 1.71 | 1 | 1 | 3.7 | 1 | 0 | 0 | 0 |
| GB Haddock | 50 | 40 | 0.75 | 1 | 1 | 2.7 | 1 | 0 | 0 | 0 |
| Witch Flounder | 45 |  | 2.45 | 3 | 2 | 7.5 | 1 | 1 | 1 | 0 |
| American Plaice | 40 | 32 | 1.70 | UNK | 1 | 4.5 | 1 | 0 | 0 | 0 |
| Pollock | 75 | 52 | 0.46 | 2 | 2 | 4.5 | 0 | 0 | 0 | 1 |
| Acadian Redfish | 30 | 25 | 0.76 | 1 | 2 | 3.8 | 1 | 1 | 0 | 0 |
| Atlantic Halibut | 45 | NA | 28.82 | UNK | 2 | 32.7 | 1 | 1 | 1 | 1 |
| Ocean Pout | 60 | NA | 12.05 | UNK | 1 | 14.9 | 0 | 1 | 1 | 1 |
| Northern (GOM-GB) <br> Windowpane Flounder | 30 | 24 | 3.48 | UNK | 2 | 7.3 | 1 | 1 | 1 | 1 |
| Southern (SNE-MA) <br> Windowpane Flounder | 30 | 24 | 0.69 | UNK | 2 | 4.5 | 1 | 1 | 1 | 1 |
| Atlantic Wolffish | 45 | NA | 3.48 | UNK | UNK | 7.0 | 1 | 0 | 0 | 0 |
| Sum |  |  |  |  |  | 174.5 | 18 | 8 | 5 | 10 |
| Mean |  |  | 5.21 | 1.83 | 1.68 | 8.73 |  |  |  |  |

[^0]The CATT also examined the suitability of sea sampling data and tagging data for this purpose as well. Sea sampling data were not suitable for this purpose because large areas are undersampled due to regulatory effects of area closures, regional catch limits, or other factors. To analyze catch distributions, the sea sampling data would further more have to be standardized with respect to vessel, gear, and possibly other factors. If not properly adjusted, clusters or hotspots using these data may have biases that identify areas where a single large vessel with large gear frequently fishes, rather than a localized high abundance or biomass of fish. Sea sampling data would also have very limited utility for analyzing distributions of groundfish due to selectivity.

Tagging data is potentially useful from two perspectives. Often, ripe and running ripe fish are identified by external examination (Figure 5). When the tag return data are adjusted for fishing effort to account for varying opportunities to catch tagged fish, the information could be useful to determine retention rates in existing or potential future closed areas. Fish that are retained for longer periods would tend to benefit more from closures than more transient fish. Unfortunately, the existing tag data tends to be relatively inaccessible (behind a Unix firewall in a foreign SQL data base), are not effort adjusted, and most tagging is done on only a few species. So the CATT felt that the tagging data had limited utility for identification of persistent spawning aggregations.

Other information was also examined or analyzed. Literature about regional groundfish spawning was examined, compiled, and taken into consideration (see Table 3and Figure 1 to Figure 5 below). Most papers were fairly general or focused on specific areas. A few, for example Ames 2004 and Deese 2005, provide broad-scale evaluation of spawning distributions, observed by fishermen. Working with Sam Truesdell at Universtiy of Maine Orono, the CATT also conducted a juvenile habitat association analysis for Gulf of Maine cod and Georges Bank cod and yellowtail flounder, applying a general additive model approach. Information from these sources was considered during the analysis and interpretation of the hotspot analysis results, but are not being reviewed in depth by the SSC.

With assistance from Owen Liu of EDF, the CATT also examined four case studies around the world where spatial management was employed in temperate fisheries that are managed with quotas. Conclusions about those studies may help influence the overall design of juvenile groundfish habitat and spawning areas.

Lastly, working with Sam Truesdell of University of Maine, Orono, the CATT developed an exploratory analysis of habitat association for three stocks: Gulf of Maine cod, Georges Bank cod, and Georges Bank yellowtail flounder. The results of this analysis were promising and for the Gulf of Maine largely corroborated the CATT's hotspot analysis for juvenile cod. A full report of this analysis is presented in a different SSC document. The results were not quantitatively used to design and propose juvenile groundfish area management options, but provided support for the options that were developed, particularly for a coastal juvenile groundfish habitat area option.

Based on the above analyses, the CATT proposed two area management options to conserve juvenile groundfish habitat. One option (Figure 14) includes all areas in the Gulf of Maine in depths less than 90 m and within 15 nm of the coastline. A second option (Figure 15) is a
network of areas that include most of the weighted hotspots from the above analysis. These area management options would be applied year round to protect vulnerable juvenile groundfish habitat, even though some groundfish species utilize the habitat on a seasonal basis.

The CATT also proposed three area management options to reduce impacts on large spawning groundfish. These management options would limit fishing activity for gears capable of catching groundfish to reduce impacts on spawning behavior and activity of large mature groundfish.

One spawning area option (Figure 16) is a network of areas that encompass the majority of the weighted hotspots. These areas would close seasonally. Areas in the Western Gulf of Maine would close following a similar seasonal progression as the existing rolling closures they would replace. A second spawning area option (Figure 17 to Figure 19) is a modification of the existing rolling closures for sector vessels, which would include all of the existing Western Gulf of Maine area and run from March to June (instead of April to June). A third option would retain a spring closure for the existing Western Gulf of Maine area and all of Closed Area II.

Table 3. Summary of groundfish spawning and habitat associations.

|  | Identified Spawning Locations | Spawning Notes | Habitat Area Location/Characteristics | Habitat Notes |
| :---: | :---: | :---: | :---: | :---: |
| Cod | Gulf of Maine: Ames Study Areas (Ames 2004). Ipswich Bay (specific spawning aggregation at Whaleback feature)(Siceloff and Howell 2012). Cape Cod Bay, western Maine coast, Jeffries Ledge and Northern Mass. Bay (Deese 2005 and Dean et al. 2012), inshore aggregations in Area 133 in the western GOM (Morin 2000) <br> Georges Bank: concentrated in the Northeast area (mostly gravel and complex relief levels)(Berlinsky 2009). | Spring spawning in northern GOM (Berlinsky 2009). <br> Fall spawning in inshore areas from Cape Cod to Nantucket Shoal (Deese 2005). <br> Winter spawning in southern GOM and Coxes Ledge (Deese 2005). <br> Spawning occurs year-round but with peaks in the summer and from Nov - Feb (Tallack 2008). <br> Spring and winter spawning in western GOM (Berlinsky 2009 and Morin 2000). <br> Peak Georges Bank spawning activity occurs in FebruaryMarch (Lough 2010) | Juveniles (age 0-1) prefer gravel substrates with lower bathymetric relief (Gregory et al. 1997) <br> Older and larger cod would move to coarse substrates with higher bathymetric relief, such as humps and ridges (Gregory et al. 1997). <br> Ipswich Bay, Mass. Bay and Cape Cod Bay (Howe et al 2002). <br> Spread across Georges Bank in early summer, constant concentration in NE Georges Bank (Lough 2010). | Age 0 cod prefer shallower depths (<90’) and move to deeper waters both in autumn and as they grow older (Howe et al. 2002) <br> Young juveniles would hide in cobble to avoid predators, and would partially remain after the threat was removed (Gotceitas and Brown, 1993). |
| Haddock | Georges Bank: Concentrated in Eastern and Northeastern areas (Overholtz 1987). | Peak spawning in Georges Bank from late March-early April (Overholtz 1987) <br> Ideal temperatures from $4-7^{\circ} \mathrm{C}$ at depths from 28-110' (Overholtz 1987) | Spread throughout Georges Bank | As pelagic juveniles grow, they move deeper in the water column (Lough and Potter 1994). |
| Yellowtail Flounder |  |  | Eastern Georges Bank, specifically within Closed Area II. (Pereira et al 2012) | Occupied area in Georges Bank doubled from $\sim 4000$ to $\sim 8000$ $\mathrm{km}^{2}$ when abundance increased (Pereira et al 2012) |


| Winter Flounder | Plymouth Bay (minor <br> activity in Plymouth <br> Estuary) (DeCelles <br> and Cadrin 2010) | Peak spawning in <br> March-May in the <br> Plymouth Bay <br> (DeCelles and Cadrin <br> 2010) |  |  |
| :--- | :--- | :--- | :--- | :--- |

## Additional figures



Figure 1. Map of indicated cod spawning areas. Circled areas indicate former spawning grounds that are no longer active. Ames, 2004.


Figure 2. Proposed cod spawning complexes. Berlinsky, 2005.


Figure 3. Summary of cod spawning areas. Deese, 2005.


Figure 4. Bathymetric map of Ipswich Bay. Black dotted rectangle highlights the elevated bathymetric feature "Whaleback". Siceloff and Howell, 2012.

NRCTP spawning fish: $\mathbf{n = 1 0 2 8}$ releases and $\mathbf{n = 5 7 0}$


Figure 5. The distribution of tagged cod releases and recaptures in spawning condition, relative to closed areas and across all years. Tallack, 2008.

Juveniles and adults were distinguished based on lengths-at-maturity for each species, which was defined according to the length at which $50 \%$ of the fish in a population mature sexually. For most species, these sizes vary by sex and stock units. They also vary over time, according to changes in growth rate, sometimes considerably. Lengths used to distinguish juveniles and adults for most species were based on data reported by O’Brien et al. (1993). Lengths at maturity for the skate species were based on information included in EFH source documents. These lengths are listed in Table 4. In most cases, O’Brien et al. based 50\% lengths at maturity on females; if there was more than one size available because of analyses that were performed at different time periods or for different stocks, they were averaged.
$r(l)=\{\exp (a+b l) /[1+\exp (a+b l)]\}$

Table 4. Lengths-at-maturity used to distinguish juveniles and adults in EFH designations. Juveniles are less than the specified length; adults are equal to or larger.

| Species | Length ( cm ) at 50\% Maturity 0 'Brien et al. (1993) and EFH Skate Source Document | Length (cm) at maturity (rounded to nearest 5 cm for analysis of juvenile and spawning distributions) <br> Calculated from parameters in latest assessment, generally GARM III <br> Red values are average L20/ L50 and L80/ L50 ratios of other species |  |  | Approximate length (rounded up to 5 cm increment) at greater than $80 \%$ Maturity from 2002-2012 spring and fall trawl survey data |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\underline{120}$ | L50 | L80 |  |
| American Plaice | 27 | 23.6 (25) | 27.6 | 31.6 (30) | 30 |
| Atlantic Cod | 35 | 35.4-36.8 (35) | 43-44.5 | 49.2-53.6 (50) | 50 |
| Atlantic Herring | 25 | (20) | NA | (25) | 25 |
| Barndoor <br> Skate | 102 | (85) | NA | (115) | 115* |
| Clearnose <br> Skate | 61 | (50) | NA | (70) |  |
| Deep-sea Red Crab | 8 |  | NA |  |  |
| Goosefish | 43 | (35) | NA | (45) | 45 |
| Haddock | 32 | 28.2-28.3 (30) | 33-34.7 | 37.8-41.1 (40) | 40 |
| Little Skate | 50 | (45) |  | (55) |  |
| Ocean Pout | 29 |  |  |  |  |
| Offshore <br> Hake | 30 | (25) |  | (35) |  |
| Pollock | 39 | 38.8 (40) | 45.4 | 51.9 (50) | 45 |
| Red Hake | 26 | (20) |  | (35) | 35 |
| Redfish | 22 | 19.2 (20) | 22.0 | 24.8 (25) | 25 |
| Rosette <br> Skate | 46 | (40) |  | (55) |  |
| Sea Scallop | 10 |  |  |  |  |
| Silver Hake | 23 | (20) |  | (30) | 30 |
| Smooth <br> Skate | 56 | (50) |  | (65) |  |
| Thorny Skate | 84 | (70) |  | (95) |  |


| Species | Length (cm) at 50\% Maturity O'Brien et al. (1993) and EFH Skate Source Document | Length (cm) at maturity (rounded to nearest 5 cm for analysis of juvenile and spawning distributions) <br> Calculated from parameters in latest assessment, generally GARM III <br> Red values are average $L 20$ / $L 50$ and $L 80$ / $L 50$ ratios of other species |  |  | Approximate length (rounded up to 5 cm increment) at greater than $80 \%$ Maturity from 2002-2012 spring and fall trawl survey data |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\underline{10}$ | L50 | L80 |  |
| White Hake | 35 | 25.0 (25) | 35.1 | 45.2 (45) | 60 |
| Windowpane | 22 | 17.5-18.2 (20) | 20.5-21.3 | 23.5-24.4 (25) |  |
| Winter Flounder | 27 | 26.7 (25) | 29-29.1 | 31.1 (30) | 30 |
| Winter Skate | 85 | (70) |  | (95) |  |
| Witch Flounder | 30 | 28.1 (30) | 32.9 | 31.1 (40) | 40 |
| Yellowtail Flounder | 27 | 24.6-25.8 (25) | 27.4-28.2 | 30.2-30.7 (30) | 30 |

Wolffish - 47 cm (Templeman 1986)

Synopsis of juvenile groundfish habitat and spawning analysis
Table 5. Cumulative proportion of abundance at age by species, survey, and stock area. First line of data represents an approximate L20 for each species. Second line of data represents a size that approximates the $90^{\text {th }}$ percentile of age 1 fish (some species use age 2 ) for the predominate stock area for each species.


Figure 6. Domain of surveys used in the hotspot analysis by season.

| Spring | Summer |
| :---: | :---: |
|  |  |
| Fall | Winter |



Figure 7. Frequency distribution plots of 2002-2012 NMFS spring trawl catches of cod <=25 cm. Top - untransformed kg/tow; Middle Catches adjusted for the proportion of zero tows in strata; Bottom - Log transformed adjusted catches.


## Field



Table 6. Cumulative number of cod caught by survey over time by size range, compared to 20 percent of total abundance.

| COMNAME REGION | ATLANTIC COD (Multiple Items) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| Row Labels $\quad$ I | 20Pct total num | Num <= 5 cm | Num $<=10 \mathrm{~cm}$ | Num $<=15 \mathrm{~cm}$ | Num <= 20 cm | Num <= 25 cm | Num <= 30 cm | Num <= 35 cm |
| ${ }^{-}$IBS Cod Spawning | g 713 | 0 | 1 | 46 | 200 | 309 | 610 | 1,340 |
| - WINTER | 353 | 0 | 1 | 31 | 99 | 128 | 270 | 137 |
| 2002-2012 | 353 | 0 | 1 | 31 | 99 | 128 | 270 | 737 |
| ${ }_{-S P R I N G}$ | 360 | 0 | 0 | 15 | 101 | 181 | 340 | 608 |
| 2002-2012 | 360 | 0 | 0 | 15 | 101 | 181 | 340 | 608 |
| ENMFS trawl | 19,013 | 1,824 | 4,110 | 6,547 | 9,888 | 14,750 | 22,563 | 32,232 |
| - WINTER | 602 | 2 | 21 | 98 | 247 | 419 | 514 | 599 |
| 1963-1971 | 314 | 1 | 20 | 32 | 61 | 118 | 159 | 210 |
| 1972-1981 | 92 | 0 | 0 | 0 | 1 | 14 | 22 | 26 |
| 1992-2001 | 153 | 1 | 1 | 6 | 34 | 94 | 132 | 162 |
| 2002-2012 | 44 | 0 | 0 | 60 | 152 | 194 | 200 | 201 |
| ESPRING | 9,157 | 1,692 | 1,815 | 2,339 | 3,983 | 6,797 | 10,455 | 14,481 |
| 1963-1971 | 359 | 62 | 71 | 83 | 104 | 132 | 169 | 259 |
| 1972-1981 | 2,301 | 443 | 524 | 612 | 901 | 1,326 | 1,837 | 2.507 |
| 1982-1991 | 1,614 | 78 | 90 | 179 | 396 | 737 | 1,113 | 1,608 |
| 1992-2001 | 607 | 109 | 115 | 141 | 232 | 323 | 427 | 563 |
| 2002-2012 | 4,276 | 999 | 1,015 | 1,324 | 2350 | 4,280 | 6,908 | 9,544 |
| ESUMMER | 1,486 | 39 | 339 | 440 | 608 | 910 | 1,355 | 1,849 |
| 1963-1971 | 474 | 9 | 18 | 37 | 87 | 118 | 192 | 287 |
| 1972-1981 | 847 | 16 | 232 | 282 | 355 | 583 | 905 | 1,236 |
| 1982-1991 | 23 | 0 | 1 | 1 | 2 | 6 | 14 | 23 |
| 1992-2001 | 142 | 14 | 88 | 120 | 159 | 208 | 244 | 302 |
| -FAll | 7,761 | 91 | 1,936 | 3,670 | 5,055 | 6,623 | 10,240 | 15,304 |
| 1963-1971 | 660 | 7 | 131 | 190 | 254 | 354 | 523 | 804 |
| 1972-1981 | 2,140 | 40 | 215 | 356 | 592 | 1,045 | 1,796 | 2877 |
| 1982-1991 | 1,111 | 18 | 180 | 299 | 426 | 675 | 1,199 | 1,847 |
| 1992-2001 | 674 | 8 | 57 | 158 | 188 | 272 | 494 | 810 |
| 2002-2012 | 3,182 | 18 | 1,352 | 2,667 | 3,593 | 4,27 | 6,228 | 8,966 |
| - MADMF trawl | 38,071 | 136,436 | 162,095 | 166,907 | 172,556 | 178,159 | 181,176 | 183,722 |
| -SPRING | 32,859 | 116,635 | 140,697 | 143,081 | 148,325 | 153,380 | 155,937 | 158,110 |
| 1972-1981 | 3,072 | 4,436 | 8,954 | 9,685 | 11,532 | 12,971 | 13,542 | 14,042 |
| 1982-1991 | 1,382 | 1,383 | 1,921 | 2,440 | 3,155 | 4,127 | 4,824 | 5,319 |
| 1992-2001 | 3,374 | 9,900 | 12,231 | 12,645 | 13,756 | 14,940 | 15,365 | 15,831 |
| 2002-2012 | 25,082 | 100,916 | 117,592 | 118,312 | 119,882 | 121,342 | 122,206 | 122.919 |
| -FALI | 5,212 | 19,801 | 21,398 | 23,826 | 24,231 | 24,779 | 25,239 | 25,612 |
| 1972-1981 | 1,874 | 7,580 | 7,716 | 8,444 | 8,544 | 8,789 | 9,006 | 9,202 |
| 1982-1991 | 1,845 | 8,230 | 8,731 | 8,892 | 8,981 | 9,102 | 9,159 | 9,201 |
| 1992-2001 | 563 | 1,124 | 1,535 | 2,668 | 2709 | 2.746 | 2766 | 2.787 |
| 2002-2012 | 929 | 2,867 | 3,416 | 3,822 | 3,996 | 4,142 | 4,308 | 4,423 |
| Grand Total | 57,796 | 138,260 | 166,206 | 173,500 | 182,644 | 193,218 | 204,348 | 217,294 |

Table 7. Cumulative weight of cod caught by survey over time by size range, compared to 20 percent of total weight.

| COMNAME REGION | ATLANTIC COD <br> (Multiple Items) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Row Labels | I 20 pct total weight | Wgt $>=50 \mathrm{~cm}$ | Wgt $>=55 \mathrm{~cm}$ | Wgt $>=60 \mathrm{~cm}$ | Wgt $>=65 \mathrm{~cm}$ | Wgt $>=70 \mathrm{~cm}$ | Wgt $>=75 \mathrm{~cm}$ | Wgt $>=80 \mathrm{~cm}$ | Wgt $>=85 \mathrm{~cm}$ | Wgt $>=90 \mathrm{~cm}$ |
| E IBS Cod Spawning | g 747 | 2.408 | 2.110 | 1,855 | 1,624 | 1,347 | 1,064 | 798 | 593 | 430 |
| - WINTER | 219 | 315 | 184 | 117 | 94 | 50 | 28 | 7 | 7 | 0 |
| 2002-2012 | 219 | 315 | 184 | 117 | 94 | 50 | 28 | 7 | 7 | 0 |
| - SPRING | 528 | 2,098 | 1,926 | 1,738 | 1,530 | 1,296 | 1,036 | 791 | 587 | 430 |
| 2002-2012 | 528 | 2,093 | 1,926 | 1,738 | 1,530 | 1,296 | 1,036 | 791 | 587 | 430 |
| E NMFS trawl | 30,250 | 126,234 | 116,874 | 105,602 | 91,915 | 78,010 | 64,149 | 52,264 | 40,615 | 31,445 |
| - WINTER | 1,654 | 7,744 | 7,421 | 6,875 | 6,273 | 5,663 | 5,002 | 4,400 | 3,594 | 2,866 |
| 1963-1971 | 1,071 | 5,112 | 4,959 | 4,720 | 4,403 | 4,013 | 3,596 | 3,173 | 2,661 | 2,128 |
| 19/2-1981 | 306 | 1,452 | 1,397 | 1,246 | 1,127 | 1,0\% | 1,010 | 923 | 717 | 632 |
| 1992-2001 | 269 | 1,159 | 1,046 | 891 | 124 | 510 | 395 | 305 | 156 | 105 |
| 2002-2012 | 8 | 21 | 18 | 18 | 18 | 9 | 92 | 0 | 0 | 0 |
| - SPRING | 14,558 | 59,891 | 55,519 | 50,284 | 43,393 | 36,609 | 29,872 | 24,347 | 18,652 | 14,300 |
| 1963-1971 | 1,141 | 5,430 | 5,229 | 4,938 | 4,517 | 4,148 | 3,620 | 3,126 | 2,501 | 1,990 |
| 19/2-1981 | 4,480 | 18,878 | 17,665 | 16,273 | 14,448 | 12,238 | 10,391 | 8,984 | 1,183 | 5,748 |
| 1982-1991 | 3,639 | 16,391 | 15,546 | 14,307 | 12,278 | 10,593 | 8,661 | 6,889 | 5,323 | 4,055 |
| 1992-2001 | 1,381 | 6,317 | 5,887 | 5,359 | 4,720 | 4,063 | 3,341 | 2,706 | 1,971 | 1,462 |
| 2002-2012 | 3,911 | 12,875 | 11,253 | 9,408 | 1,430 | 5,567 | 3,860 | 2,642 | 1,668 | 1,047 |
| -SUMMER | 2889 | 12,728 | 11,587 | 10,206 | 8,948 | 7,575 | 6,234 | 4,992 | 3,984 | 3,132 |
| 1963-19/1 | 1,201 | 5,566 | 5,241 | 4,789 | 4,186 | 3,500 | 2,851 | 2,317 | 1,769 | 1,329 |
| 19/2-1981 | 1,455 | 6,301 | 5,544 | 4,735 | 4,162 | 3,498 | 2,915 | 2,279 | 1,897 | 1,557 |
| 1982-1991 | 42 | 172 | 147 | 132 | 104 | 83 | 72 | 68 | 51 | 26 |
| 1992-2001 | $1 / 4$ | 689 | 635 | 550 | 496 | 444 | 395 | 328 | 261 | 220 |
| FFALL | 11,158 | 45,872 | 42,307 | 38,236 | 33,302 | 28,213 | 23,040 | 18,526 | 14,445 | 11,145 |
| 1963-1971 | 1,684 | 1,793 | 7,458 | 6,993 | 6,330 | 5,665 | 4,982 | 4,275 | 3,540 | 2,821 |
| 19/2-1981 | 4,366 | 19,429 | 18,092 | 16,496 | 14,560 | 12,593 | 10,480 | 8,6/8 | 1,0/3 | 5,590\% |
| 1982-1991 | 1,6/9 | 6,914 | 6,397 | 5,710 | 4,8/9 | 3,990 | 3,271 | 2,553 | 1,888 | 1,493 |
| 1992-2001 | 1,063 | 4,411 | 3,899 | 3,322 | 2,717 | 2,131 | 1,512 | 1,019 | 702 | 431 |
| 2002-2012 | 2,365 | 1,325 | 6,461 | 5,716 | 4,816 | 3,834 | 2,789 | 2,001 | 1,242 | 810 |
| EMADMF trawl | 2,206 | 5,354 | 4,219 | 3,313 | 2,459 | 1,761 | 1,129 | 736 | 546 | 409 |
| - SPRING | 2.038 | 5,097 | 4,015 | 3,140 | 2.330 | 1,681 | 1,090 | 715 | 533 | 404 |
| 1972-1981 | 401 | 836 | 621 | 445 | 291 | 208 | 149 | 110 | 87 | 71 |
| 1982-1991 | 414 | 142 | 533 | 369 | 264 | 180 | 148 | 122 | 101 | 600 |
| 1992-2001 | 320 | 633 | 475 | 347 | 225 | 155 | 105 | 60 | 35 | 20 |
| 2002-2012 | $8 \%$ | 2,886 | 2,381 | 1,9880 | 1,544 | 1,138 | 6888 | 423 | 310 | 252 |
| - FAll | 168 | 251 | 204 | 173 | 130 | 86 | 39 | 22 | 13 | 5 |
| 1972-1981 | 61 | 53 | 44 | 37 | 21 | 25 | 16 | 13 | 13 | 5 |
| 1982-1991 | 16 | 11 | 9 | 1 | 4 | 3 | 0 | 0 | 0 | 0 |
| 1992-2001 | 13 | 12 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002-2012 | 78 | 182 | 147 | 126 | 98 | 51 | 23 | 9 | 0 | 0 |
| Grand Total | 33,202 | 133,997 | 123,202 | 110,770 | 95,998 | 81,124 | 66,342 | 53,798 | 41,814 | 32,284 |

Synopsis of juvenile groundfish habitat and spawning analysis
Table 8. Cumulative biomass above 5 cm size ranges by species, survey, and decade, compared to $20 \%$ of total weight per tow (kg) and the size at estimated $80 \%$ maturity for females.

| Approximate 20\% of biomass (upper), 180 for maturity (lower) | Species | T Row Labels |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 75 cm | athanticcod | All | 30,250 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| L80 $=50 \mathrm{~cm}$ | athanticcod | winter | 1,654 | 8,247 | 8,226 | 8,202 | 8,141 | 7,983 | 7,744 | 7,421 | 6,875 | 6,273 | 5,663 | 5,002 | 4,400 | 3,594 | 2,866 | 1,978 | 1,353 |
|  | ATLANTICCOD | 1963-1971 | 1,071 | 5,348 | 5,339 | 5,325 | 5,291 | 5,222 | 5,112 | 4,959 | 4,720 | 4,403 | 4.013 | 3,596 | 3,173 | 2,661 | 2,128 | 1,461 | 1,016 |
|  | ATLANTICCOD | 1972-1981 | 306 | 1.530 | 1.528 | 1,527 | 1,517 | 1,488 | 1,452 | 1,397 | 1,246 | 1,127 | 1,070 | 1,010 | 923 | 77 | 632 | 460 | 312 |
|  | ATAANTICCOD | 1992-2001 | 269 | 1,339 | 1,330 | 1,321 | 1,305 | 1,247 | 1,159 | 1,046 | 891 | 724 | 570 | 395 | 305 | 156 | 105 | 57 | 25 |
|  | ATLANTICCOD | 2002-2012 |  | 30 | 29 | 29 | 28 | 26 |  | 18 | 18 | 18 | 9 | 2 | 0 | 0 | 0 |  | 0 |
|  | ATAANTICCOD | SPRING | 14,558 | 72,457 | 71,801 | 70,561 | 68,244 | 64,198 | 59,891 | 55,579 | 50,284 | 43,393 | 36,609 | 29,872 | 24,347 | 18,652 | 14,302 | 10,866 | 7,891 |
|  | ATLANTICCOD | $1963-1971$ | 1,141 | 5,701 | 5,966 | 5,672 | 5,614 | 5,551 | 5,430 | 5.229 | 4,938 | 4,517 | 4.148 | 3.620 | 3,126 | 2,501 | 1,990 | 1,516 | 1,130 |
|  | ATLANTICCOD | 1972-1981 | 4,480 | 22,342 | 22,248 | 22,062 | 21,645 | 20,446 | 18.878 | 17,665 | 16,273 | 14,448 | 12,238 | 10,391 | 8,984 | 7,183 | 5,748 | 4,489 | 3,320 |
|  | ATLANTICCOD | 1982-1991 | 3.639 | 18,153 | 18,082 | 17,935 | 17,643 | 17,118 | 16,391 | 15.546 | 14,307 | 12,278 | 10,593 | 8.661 | 6.889 | 5,323 | 4,055 | 3,222 | 2,343 |
|  | ATANTICCOD | 1992-2001 | 1,387 | 6,923 | 6,906 | 6,864 | 6,778 | 6,591 | 6,317 | 5,887 | 5,359 | 4,720 | 4,063 | 3,341 | 2,706 | 1,977 | 1,462 | 1,007 | 675 |
|  | athantic cod | 2002-2012 | 3,911 | 19,338 | 18.869 | 18.028 | 16.564 | 14,492 | 12,875 | 11,253 | 9.408 | 7,430 | 5.567 | 3.860 | 2,642 | 1,668 | 1,047 | 632 | 423 |
|  | ATAANTICCOD | SUMMER | 2,879 | 14,357 | 14,282 | 14,124 | 13,863 | 13,478 | 12,728 | 11,567 | 10,206 | 8,948 | 7,525 | 6,334 | 4,992 | 3,984 | 3,132 | 2,334 | 1,736 |
|  | ATLANTICCOD | 1963 -1971 | 1,207 | 6,032 | 6.020 | 5.991 | 5,927 | 5.799 | 5,566 | 5,241 | 4,789 | 4,186 | 3,500 | 2,851 | 2,317 | 1,769 | 1,329 | 974 | 726 |
|  | ATAANTICCOD | 1972-1981 | 1,455 | 7,252 | 7,197 | 7,088 | 6,936 | 6,745 | 6,301 | 5.544 | 4,735 | 4,162 | 3,498 | 2,915 | 2,279 | 1,897 | 1,557 | 1,169 | 874 |
|  | ATAANTICCOD | 1982-1991 | 42 | 209 | 207 | 205 | 203 | 195 | 172 | 147 | 132 | 104 | 83 | 72 | 68 | 51 | 26 | 26 |  |
|  | ATAANTICCOD | 1992-2001 | 174 | 864 | 858 | 840 | 796 | 739 | 689 | 635 | 550 | 496 | 444 | 395 | 328 | 267 | 220 | 166 | 111 |
|  | ATAANTICCOD | Fall | 11,158 | 55,545 | 54,962 | 53,397 | 50,972 | 48,454 | 45,872 | 42,307 | 38,236 | 33,302 | 28,213 | 23,40 | 18,526 | 14,455 | 11,145 | 8,424 | 6,170 |
|  | ATAANTICCOD | 1963-1971 | 1,684 | 8.407 | 8,379 | 8,292 | 8,177 | 8.005 | 7,793 | 7,458 | 6,993 | 6,330 | 5,665 | 4,982 | 4,275 | 3,540 | 2,821 | 2,220 | 1,622 |
|  | ATLANTICCOD | 1972-1981 | 4,366 | 21,777 | 21,653 | 21,317 | 20,808 | 20,197 | 19,429 | 18.092 | 16,496 | 14,560 | 12,593 | 10,480 | 8.678 | 7.073 | 5,590 | 4,351 | 3,324 |
|  | ATLANTICCOD | 1982-1991 | 1,679 | 8,367 | 8,280 | 8.078 | 7.697 | 7,259 | 6,914 | 6,397 | 5,710 | 4.879 | 3.990 | 3,277 | 2,553 | 1,888 | 1,493 | 1,046 | 724 |
|  | ATANTICCOD | 1992-2001 | 1,063 | 5,306 | 5,269 | 5,173 | 4,995 | 4,742 | 4,411 | 3,899 | 3,322 | 2,717 | 2,131 | 1,512 | 1,019 | 702 | 431 | 293 | 156 |
|  | ATANTICCOD | 2002-2012 | 2,365 | 11,688 | 11,380 | 10,536 | 9,295 | 8,252 | 7,325 | 6,461 | 5,716 | 4,816 | 3,834 | 2,789 | 2,001 | 1,242 | 810 | 514 | 344 |
| 40 cm | american plaice | winter | 62 | 310 | 300 | 261 | 202 | 130 | 76 | 47 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 $=30 \mathrm{~cm}$ | american plaice | 1972-1981 | 17 | 85 | 83 | 75 | 63 | 41 | 32 | 27 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | AMERICAN PLIICE | 1992-2001 | 44 | 219 | 212 | 182 | 136 | 88 | 44 | 21 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLICE | 2002-2012 | 1 | 6 | 5 | 5 | 4 | 1 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , |
|  | AMERICAN PLIICE | SPRING | 2,992 | 11,176 | 9,366 | 6,995 | 4,939 | 3,250 | 1,793 | 763 | 289 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | AMERICAN PLIICE | 1963-1971 | 233 | 1,113 | 972 | 756 | 543 | 359 | 194 | 109 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLIICE | 1972-1981 | 1,076 | 4,968 | 4,453 | 3,662 | 2,815 | 1,951 | 1,089 | 482 | 167 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | AMERICAN PLIICE | 1982-1991 | 453 | 2,007 | 1,647 | 1,216 | 861 | 601 | 366 | 137 | 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLIICE | 1992-2001 | 338 | 1,498 | 1,173 | 157 | 457 | 234 | 105 | 33 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | , |  |
|  | AMERICAN PLICE | 2002-2012 | 392 | 1,589 | 1,122 | 603 | 264 | 106 | 38 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | AMERICAN PLIICE | SUMMER | 924 | 4,013 | 3,153 | 2,062 | 1,264 | 793 | 424 | 171 | 62 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLICE | 1963-1971 | 81 | 385 | 331 | 244 | 172 | 104 | 65 | 36 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | AMERICAN PLICE | 1972-1981 | 434 | 1,875 | 1,556 | 1,196 | 835 | 544 | 296 | 125 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | AMERICAN PLIICE | 1982-1991 | 81 | 350 | 216 | 73 | 20 | 11 | 6 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLICE | 1992-2001 | 328 | 1,402 | 1,049 | 549 | 237 | 134 | 57 | 11 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | American plaice | Fall | 2,690 | 12,037 | 10,086 | 7,423 | 5,086 | 3,152 | 1,750 | 768 | 244 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | AMERICAN PLICE | 1963-1971 | 171 | 812 | 706 | 540 | 368 | 224 | 138 | 79 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | american plaice | 1972-1981 | 1,248 | 5,780 | 5,148 | 4,197 | 3,186 | 2.113 | 1,221 | 535 | 169 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | am Erican plaice | 1982-1991 | 412 | 1,777 | 1,418 | 982 | 673 | 422 | 234 | 103 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | AMERICAN PLICE | 1992-2001 | 504 | 2,217 | 1,785 | 1,119 | 578 | 265 | 109 | 33 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | AMERICAN PLICE | 2002-2012 | 355 | 1,452 | 1,030 | 586 | 281 | 128 | 48 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | AMERICAN PLICE | All | 6,168 | 27,535 | 22,904 | 16,741 | 11,991 | 7,327 | 4,042 | 1,750 | 617 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | ATAANTC HERRING | WINTER | 304 | 765 | 85 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 $=\mathbf{2 5} \mathrm{cm}$ | atantic herring | 1963-1971 | 8 | 23 | 3 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | ATAANTC Herring | 1972-1981 | 9 | 22 | 3 |  | , | 0 |  | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | ATAANTC HERRING | 1992-2001 | 260 | 670 | 77 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIING | 2002-2012 |  | 49 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | ATAANTC Herring | SPRING | 2,253 | 4,363 | 255 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATtantic herring | 1963-1971 | 10 | 23 | 9 | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATAANTC HERRIING | 1972-1981 | 239 | 649 | 83 | 2 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATAANTC HERRING | 1982-1991 | 321 | 1,063 | 104 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATAANTC HERRING | 1992-2001 | 778 | 1,738 | 46 | 1 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIIN | 2002-2012 | 906 | 890 |  | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , |  |
|  | ATLANTC HERRIIN | SUMMER | 1,782 | 5,508 | 927 | 69 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATAANTC HerRing | 1963-1971 | 229 | 1,088 | 615 | 68 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIIN | 1972-1981 | 64 | 220 | 37 | 0 | , | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 |
|  | ATLANTC HERRIING | 1982-1991 | 484 | 1.224 | 112 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTIC HERRING | 1992-2001 | 1,006 | 2,976 | 164 |  | 1 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
|  | ATLANTC HERRIIN | FALL | 4,896 | 12,628 | 1,070 | 6 | 0 | 0 | 0 | - | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIIN | 1963-1971 | 71 | 318 | 99 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Atantic herring | 1972-1981 | 32 | 148 | 57 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIIN | 1988-1991 | 651 | 2,285 | 513 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIING | 1992-2001 | 1,713 | 5,766 | 368 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATLANTC HERRIING | 2002-2012 | 2,429 | 4,112 | 34 |  |  | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | ATIANTC Herring | All | 9,235 | 23,264 | 2,337 | 83 | 4 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 |

Synopsis of juvenile groundfish habitat and spawning analysis


Synopsis of juvenile groundfish habitat and spawning analysis

| Approximate 20\% of biomass (upper), 180 for maturity (lower) | Species | Row Labels |  |  | $5^{0^{101}}$ | $5^{65^{1010}}$ | $5^{5^{4^{0^{8}}}}=$ | $55^{55^{5}}$ | $50^{5}$ | $5^{85^{10}}=$ | $8^{8}$ |  | $5^{4^{5}}$ | $5^{5^{44^{4}}}$ | $5^{5^{0.1}}$ | $10^{60}$ |  | $0^{65}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 50 cm | UTILE SKATE | WINTER | 4,589 | 22,768 | 22,311 | 21,183 | 19,260 | 13,916 | 2,149 | 124 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 = 55 cm | UTLLE SKATE | 1963-1971 | 457 | 2,285 | 2,281 | 2,257 | 2,170 | 1,624 | 277 | 32 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 1972-1981 | 144 | 707 | 688 | 637 | 574 | 482 | 221 | 83 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 1992-2001 | 2,721 | 13,488 | 13,186 | 12,366 | 11,071 | 7,779 | 1,152 | , | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 2002-2012 | 1,266 | 6,288 | 6,156 | 5,923 | 5,444 | 4,031 | 498 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTIE SKATE | SPRING | 4,842 | 23,84 | 23,220 | 22,036 | 20,462 | 16,028 | 3,493 | 178 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 1963-1971 | 297 | 1,476 | 1,459 | 1,424 | 1,360 | 1,104 | 239 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 1972-1981 | 1,399 | 6,915 | 6,758 | 6,428 | 5,958 | 4,685 | 1,034 | 74 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 1982-1991 | 1,088 | 5,359 | 5,205 | 4,978 | 4,665 | 3,583 | 795 | 36 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTLE SKATE | 1992-2001 | 872 | 4,277 | 4,112 | 3,858 | 3,554 | 2,752 | 604 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 2002-2012 | 1,187 | 5,857 | 5,686 | 5,349 | 4,925 | 3,905 | 820 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | SUMMER | 506 | 2,519 | 2,505 | 2,478 | 2,405 | 2,005 | 487 | 53 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 1963-1971 | 191 | 951 | 949 | 942 | 918 | 720 | 132 | 30 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTIE SKATE | 1972-1981 | 271 | 1,348 | 1,338 | 1,320 | 1,279 | 1,101 | 231 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 1982-1991 | 0 | 2 | 2 | 2 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 1992-2001 | 44 | 218 | 217 | 214 | 206 | 182 | 123 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | fall | 4,375 | 21,686 | 21,347 | 20,638 | 19,327 | 15,447 | 3,816 | 213 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTIE SKATE | 1963-1971 | 342 | 1,708 | 1,696 | 1,666 | 1,603 | 1,298 | 285 | 41 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | 1972-1981 | 1,383 | 6,853 | 6,764 | 6,598 | 6,256 | 5,192 | 1,308 | 80 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTIE SKATE | 1982-1991 | 859 | 4,242 | 4,137 | 3,927 | 3,547 | 2,701 | 727 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTIE SKATE | 1992-2001 | 940 | 4,668 | 4,604 | 4,477 | 4,255 | 3,403 | 829 | 39 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTEE SKATE | 2002-2012 | 851 | 4,215 | 4,145 | 3,970 | 3,666 | 2,853 | 666 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | UTTE SKATE | All | 14,312 | 70,856 | 69,383 | 66,335 | 61,454 | 47,397 | 9,944 | 568 | 73 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 60 cm | ocean pout | WINTER | 1,476 | 7,370 | 7,359 | 7,310 | 7,176 | 6,915 | 6,414 | 5,599 | 4,314 | 2,888 | 1,919 | 1,135 | 584 | 213 | 81 | 0 | 0 |
| NA | OCEAN POUT | 1963-1971 | 540 | 2,700 | 2,699 | 2,696 | 2,689 | 2,672 | 2,615 | 2,459 | 2,124 | 1,622 | 1,219 | 813 | 454 | 177 | 63 | 0 | 0 |
|  | OCEAN POUT | 1972-1981 | 41 | 203 | 203 | 202 | 200 | 199 | 191 | 168 | 154 | 125 | 83 | 46 | 24 | 8 | 4 | 0 | 0 |
|  | OCEAN POUT | 1992-2001 | 848 | 4,235 | 4,225 | 4,181 | 4,056 | 3,823 | 3,416 | 2,805 | 1,909 | 1,076 | 575 | 257 | 99 | 29 | 14 | 0 | 0 |
|  | OCEAN POUT | 2002-2012 | 46 | 232 | 232 | 232 | 231 | 221 | 192 | 166 | 126 | 65 | 41 | 20 | 6 | 0 | 0 | 0 | 0 |
|  | OCEAN POUT | SPRING | 2,483 | 12,390 | 12,343 | 12,201 | 11,861 | 11,029 | 9,865 | 8,242 | 6,549 | 4,631 | 3,047 | 1,720 | 904 | 381 | 137 | 0 | 0 |
|  | OCEAN POUT | 1963-1971 | 146 | 728 | 728 | 725 | 718 | 684 | 607 | 549 | 467 | 370 | 283 | 159 | 94 | 41 | 24 | 0 | 0 |
|  | ocean pout | 1972-1981 | 710 | 3,541 | 3,527 | 3,484 | 3,363 | 2,974 | 2,517 | 2,010 | 1,575 | 1,128 | 743 | 455 | 281 | 125 | 43 | 0 | 0 |
|  | OCEAN POUT | 1982-1991 | 1,111 | 5,546 | 5,529 | 5,473 | 5,343 | 5,078 | 4,685 | 3,986 | 3,196 | 2,271 | 1,468 | 829 | 410 | 175 | 67 | 0 | 0 |
|  | OCEAN POUT | 1992-2001 | 353 | 1,764 | 1,759 | 1,742 | 1,706 | 1,621 | 1,471 | 1,209 | 914 | 598 | 392 | 201 | 89 | 33 | 3 | 0 | 0 |
|  | OCEAN POUT | 2002-2012 | 163 | 810 | 801 | 776 | 732 | 671 | 585 | 489 | 397 | 264 | 162 | 76 | 31 | 6 | 0 | 0 | 0 |
|  | OCEAN POUT | SUMMER | 277 | 1,384 | 1,375 | 1,345 | 1,277 | 1,170 | 1,042 | 918 | 787 | 629 | 453 | 273 | 146 | 55 | 26 | 0 | 0 |
|  | OCEAN POUT | 1963-1971 | 95 | 473 | 472 | 471 | 466 | 459 | 452 | 439 | 407 | 340 | 236 | 128 | 62 | 28 | 12 | 0 | 0 |
|  | OCEAN POUT | 1972-1981 | 127 | 631 | 625 | 608 | 578 | 531 | 456 | 396 | 329 | 269 | 203 | 143 | 84 | 28 | 13 | 0 | 0 |
|  | OCEAN POUT | 1982-1991 | 15 | 73 | 72 | 70 | 62 | 46 | 32 | 22 | 13 | 10 | 8 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | ocean pout | 1992-2001 | 42 | 207 | 205 | 197 | 171 | 134 | 101 | 62 | 38 | 10 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | OCEAN POUT | FALL | 446 | 2,216 | 2,188 | 2,088 | 1,908 | 1,663 | 1,358 | 1,027 | 729 | 481 | 293 | 183 | 114 | 59 | 28 | 0 | 0 |
|  | OCEAN POUT | 1963-1971 | 54 | 271 | 269 | 264 | 251 | 231 | 205 | 166 | 137 | 104 | 60 | 38 | 25 | 11 | 11 | 0 | 0 |
|  | ocean pout | 1972-1981 | 151 | 752 | 744 | 725 | 686 | 620 | 526 | 404 | 291 | 185 | 137 | 97 | 63 | 40 | 13 | 0 | 0 |
|  | ocean pout | 1982-1991 | 85 | 422 | 416 | 395 | 364 | 315 | 243 | 182 | 119 | 77 | 49 | 23 | 13 | 4 | 4 | 0 | 0 |
|  | OCEAN POUT | 1992-2001 | 111 | 552 | 546 | 523 | 465 | 395 | 312 | 233 | 158 | 102 | 45 | 25 | 13 | 4 | 0 | 0 | 0 |
|  | OCEAN POUT | 2002-2012 | 45 | 219 | 212 | 182 | 142 | 102 | 72 | 42 | 25 | 14 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | OCEAN POUT | All | 4,682 | 23,360 | 23,265 | 22,943 | 22,221 | 20,777 | 18,679 | 15,786 | 12,378 | 8,629 | 5,712 | 3,311 | 1,748 | 707 | 273 | 0 | 0 |
| 75 cm | Рошоск | WINTER | 621 | 3,094 | 3,071 | 3,039 | 2,934 | 2,838 | 2,712 | 2,576 | 2,384 | 2,143 | 1,800 | 1,466 | 1,051 | 607 | 311 | 139 | 0 |
| L80 $=50 \mathrm{~cm}$ | роцоск | 1963-1971 | 505 | 2,518 | 2,495 | 2,463 | 2,359 | 2,266 | 2,142 | 2,013 | 1,845 | 1,630 | 1,351 | 1,094 | 761 | 416 | 195 | 89 | 0 |
|  | Ропоск | 1972-1981 | 106 | 529 | 529 | 528 | 528 | 525 | 523 | 517 | 498 | 473 | 413 | 340 | 273 | 174 | 105 | 40 | 0 |
|  | Pошоск | 1992-2001 | 10 | 48 | 48 | 48 | 47 | 47 | 47 | 45 | 41 | 39 | 36 | 32 | 17 | 17 | 10 | 10 | 0 |
|  | Pошоск | 2002-2012 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Pошоск | SPRING | 5,183 | 25,770 | 25,582 | 25,096 | 24,484 | 23,329 | 22,026 | 20,190 | 17,838 | 15,673 | 13,483 | 11,170 | 8,798 | 6,597 | 4,321 | 2,219 | 0 |
|  | Pошосk | 1963-1971 | 459 | 2,286 | 2,280 | 2,270 | 2,257 | 2,233 | 2,194 | 2,158 | 2,077 | 1,996 | 1,964 | 1,859 | 1,608 | 1,166 | 632 | 244 | 0 |
|  | Pошоск | 1972-1981 | 1,753 | 8,743 | 8,651 | 8,337 | 8,009 | 7,547 | 7,201 | 6,720 | 6,088 | 5,590 | 5,054 | 4,547 | 3,889 | 3,065 | 1,997 | 1,040 | 0 |
|  | PошоСК | 1982-1991 | 1,630 | 8,125 | 8,093 | 8,038 | 7,951 | 7,600 | 6,981 | 6,114 | 5,196 | 4,457 | 3,650 | 2,950 | 2,385 | 1,964 | 1,481 | 851 | 0 |
|  | Рошоск | 1992-2001 | 513 | 2,533 | 2,500 | 2,448 | 2,305 | 2,036 | 1,818 | 1,589 | 1,351 | 1,079 | 864 | 643 | 377 | 170 | 100 | 44 | 0 |
|  | Pошоск | 2002-2012 | 828 | 4,084 | 4,058 | 4,003 | 3,961 | 3,914 | 3,833 | 3,609 | 3,126 | 2,551 | 1,951 | 1,171 | 540 | 232 | 110 | 40 | 0 |
|  | Pошоск | SUMMER | 812 | 3,975 | 3,913 | 3,881 | 3,805 | 3,705 | 3,616 | 3,459 | 3,285 | 3,089 | 2,738 | 2,273 | 1,797 | 1,298 | 820 | 458 | 0 |
|  | Рошоск | 1963-1971 | 349 | 1,747 | 1,746 | 1,735 | 1,694 | 1,614 | 1,538 | 1,427 | 1,343 | 1,244 | 1,093 | 847 | 575 | 304 | 132 | 48 | 0 |
|  | Рошоск | 1972-1981 | 429 | 2,076 | 2,025 | 2,012 | 1,982 | 1,964 | 1,950 | 1,909 | 1,827 | 1,745 | 1,578 | 1,395 | 1,204 | 976 | 677 | 399 | 0 |
|  | Pошосk | 1982-1991 | 1 | 5 | 4 | 4 | 3 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Pоцоск | 1992-2001 | 33 | 147 | 138 | 131 | 126 | 125 | 125 | 121 | 115 | 100 | 67 | 32 | 19 | 19 | 11 | 11 | 0 |
|  | Роцоск | FAL | 4,206 | 20,989 | 20,736 | 20,392 | 19,826 | 18,807 | 17,416 | 15,918 | 14,777 | 13,520 | 11,736 | 9,743 | 7,499 | 5,375 | 3,642 | 2,017 | 0 |
|  | Pошоск | 1963-1971 | 681 | 3,404 | 3,400 | 3,378 | 3,319 | 3,158 | 2,965 | 2,864 | 2,780 | 2,646 | 2,318 | 1,837 | 1,256 | 794 | 504 | 285 | 0 |
|  | Pошоск | 1972-1981 | 1,975 | 9,874 | 9,845 | 9,803 | 9,614 | 9,158 | 8,848 | 8,506 | 8,104 | 7,553 | 6,771 | 5,849 | 4,797 | 3,631 | 2,526 | 1,376 | 0 |
|  | Pоцоск | 1982-1991 | 489 | 2,434 | 2,393 | 2,342 | 2,260 | 2,169 | 1,975, | 1,706. | 1,528 | 1,414 | 1,274 | 1,105 | 884 | 673 | 446 | 266 | 0 |
|  | Рошоск | 1992-2001 | 321 | 1,582 | 1,501 | 1,373 | 1,246 | 1.120 | 22 23 | Of 69\% | 578 | 462 | 323 | 199 | 143 | 88 | 54 | 31 | 0 |
|  | PошоСК | 2002-2012 | 741 | 3,694 | 3,597 | 3,497 | 3,387 | 3,202 | 2,703 | 2,092 | 1,786 | 1,446 | 1,050 | 754 | 419 | 188 | 112 | 59 | 0 |
|  | POLOCK | All | 10,822 | 53,828 | 53,302 | 52,408 | 51,049 | 48,678 | 45,770 | 42,143 | 38,284 | 34,426 | 29,757 | 24,652 | 19,145 | 13,877 | 9,095 | 4,834 | 0 |

Synopsis of juvenile groundfish habitat and spawning analysis


Synopsis of juvenile groundfish habitat and spawning analysis

| Approximate 20\% of biomass (upper), 180 for maturity (lower) | Species | R Row Labels | $=0.0^{00^{5}}$ |  |  |  | $5^{5^{4^{0^{0}}}}=$ |  | $s^{505}$ |  |  | $5^{4^{4^{8}}}$ |  | $5^{\frac{140^{4}}{40}}$ | $5^{00^{40}}$ |  | $5^{05^{10}}$ | $0^{65}$ | $55^{50^{5}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 30 cm | SILVER Hake | WINTER | 530 | 1,815 | 675 | 312 | 134 | 78 | 44 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 = 30 cm | SILVER Hake | 1963-1971 | 208 | 775 | 443 | 241 | 108 | 64 | 40 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1972-1981 | 4 | 19 | 15 | 9 | 7 | 6 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1992-2001 | 280 | 919 | 185 | 51 | 17 | 6 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER Hake | 2002-2012 | 39 | 102 | 33 | 11 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | SPRING | 3,994 | 12,959 | 6,550 | 2,564 | 1,024 | 508 | 284 | 152 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1963-1971 | 70 | 298 | 189 | 102 | 49 | 26 | 8 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1972-1981 | 1,714 | 6,911 | 4,682 | 1,876 | 727 | 381 | 219 | 115 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1982-1991 | 484 | 1,678 | 789 | 289 | 118 | 52 | 30 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1992-2001 | 1,045 | 2,517 | 486 | 183 | 90 | 33 | 20 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 2002-2012 | 681 | 1,555 | 404 | 114 | 40 | 16 | 6 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | SUMMER | 1,639 | 5,840 | 3,990 | 1,837 | 853 | 467 | 277 | 125 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1963-1971 | 571 | 2,651 | 1,873 | 821 | 354 | 184 | 114 | 50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SIIVER HAKE | 1972-1981 | 438 | 1,927 | 1,579 | 807 | 414 | 242 | 135 | 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1982-1991 | 94 | 206 | 108 | 42 | 9 | 7 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1992-2001 | 535 | 1,056 | 430 | 167 | 75 | 34 | 24 | 11 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | Fall | 6,532 | 23,582 | 13,035 | 5,751 | 2,586 | 1,322 | 727 | 364 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1963-1971 | 569 | 2,436 | 1,754 | 911 | 528 | 339 | 198 | 94 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1972-1981 | 1,417 | 6,111 | 4,801 | 2,432 | 1,091 | 630 | 401 | 222 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1982-1991 | 1,525 | 6,284 | 3,577 | 1,470 | 577 | 189 | 55 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 1992-2001 | 1,530 | 4,656 | 1,738 | 554 | 243 | 112 | 46 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | 2002-2012 | 1,491 | 4,093 | 1,167 | 384 | 148 | 53 | 27 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SILVER HAKE | All | 12,695 | 44,196 | 24,250 | 10,463 | 4,597 | 2,376 | 1,332 | 654 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 55 cm | SMOOTH SKATE | WINTER | 33 | 165 | 162 | 154 | 142 | 128 | 109 | 67 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 = 65 cm | SMOOTH SKATE | 1963-1971 | 16 | 78 | 76 | 72 | 66 | 60 | 52 | 29 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1972-1981 | 10 | 52 | 50 | 47 | 43 | 39 | 34 | 24 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1992-2001 | 7 | 35 | 35 | 34 | 33 | 29 | 23 | 14 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 2002-2012 | 0 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | SPRING | 226 | 1,115 | 1,095 | 1,057 | 995 | 900 | 712 | 382 | 109 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1963-1971 | 23 | 116 | 115 | 113 | 108 | 103 | 91 | 54 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1972-1981 | 77 | 382 | 376 | 365 | 344 | 309 | 250 | 141 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
|  | SMOOTH SKATE | 1982-1991 | 35 | 172 | 169 | 165 | 159 | 149 | 127 | 74 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1992-2001 | 25 | 124 | 122 | 116 | 112 | 102 | 75 | 36 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 2002-2012 | 66 | 322 | 313 | 298 | 272 | 236 | 168 | 76 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | SUMMER | 26 | 129 | 127 | 124 | 118 | 107 | 90 | 55 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1963-1971 | 12 | 58 | 58 | 57 | 56 | 51 | 42 | 26 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1972-1981 | 5 | 27 | 27 | 26 | 25 | 21 | 18 | 10 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1982-1991 | 2 | 12 | 11 | 11 | 9 | 9 | 8 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1992-2001 | 7 | 32 | 31 | 30 | 28 | 26 | 22 | 15 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | FALL | 247 | 1,219 | 1,199 | 1,166 | 1,118 | 1,041 | 892 | 511 | 152 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1963-1971 | 39 | 191 | 188 | 182 | 173 | 162 | 141 | 82 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1972-1981 | 58 | 291 | 289 | 285 | 278 | 261 | 223 | 124 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1982-1991 | 39 | 195 | 192 | 189 | 182 | 173 | 154 | 97 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 1992-2001 | 55 | 272 | 266 | 257 | 246 | 223 | 187 | 104 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | 2002-2012 | 56 | 271 | 264 | 253 | 240 | 222 | 188 | 105 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | SMOOTH SKATE | All | 532 | 2,628 | 2,583 | 2,502 | 2,373 | 2,176 | 1,804 | 1,015 | 296 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 85 cm | THORNY SKATE | WINTER | 592 | 2,945 | 2,927 | 2,893 | 2,852 | 2,795 | 2,723 | 2,614 | 2,482 | 2,320 | 2,130 | 1,920 | 1,640 | 1,205 | 854 | 468 | 189 |
| L80 $=95 \mathrm{~cm}$ | THORNY SKATE | 1963-1971 | 486 | 2,422 | 2,410 | 2,389 | 2,368 | 2,334 | 2,291 | 2,218 | 2,123 | 2,005 | 1,864 | 1,685 | 1,467 | 1,130 | 829 | 450 | 189 |
|  | THORNY SKATE | 1972-1981 | 83 | 413 | 409 | 404 | 395 | 382 | 362 | 339 | 313 | 280 | 243 | 215 | 158 | 69 | 25 | 18 | 0 |
|  | THORNY SKATE | 1992-2001 | 22 | 109 | 107 | 98 | 87 | 76 | 69 | 56 | 46 | 35 | 23 | 20 | 16 | 6 | 0 | 0 | 0 |
|  | THORNY SKATE | 2002-2012 | , | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | THORNY SKATE | SPRING | 2,268 | 11,258 | 11,162 | 11,035 | 10,829 | 10,557 | 10,115 | 9,495 | 8,737 | 7,931 | 7,090 | 6,159 | 5,186 | 4,047 | 2,771 | 1,691 | 869 |
|  | THORNY SKATE | 1963-1971 | 474 | 2,354 | 2,338 | 2,324 | 2,295 | 2,250 | 2,166 | 2,094 | 1,979 | 1,871 | 1,710 | 1,556 | 1,371 | 1,094 | 779 | 494 | 290 |
|  | THORNY SKATE | 1972-1981 | 1,059 | 5,262 | 5,223 | 5,162 | 5,068 | 4,944 | 4,757 | 4,448 | 4,088 | 3,683 | 3,288 | 2,801 | 2,353 | 1,914 | 1,280 | 833 | 450 |
|  | THORNY SKATE | 1982-1991 | 495 | 2,459 | 2,435 | 2,406 | 2,355 | 2,297 | 2,207 | 2,057 | 1,881 | 1,660 | 1,460 | 1,256 | 1,013 | 721 | 508 | 279 | 96 |
|  | THORNY SKATE | 1992-2001 | 134 | 663 | 654 | 643 | 625 | 599 | 556 | 510 | 446 | 397 | 353 | 309 | 254 | 168 | 103 | 41 | 22 |
|  | THORNY SKATE | 2002-2012 | 105 | 520 | 512 | 501 | 486 | 468 | 429 | 385 | 344 | 320 | 279 | 237 | 195 | 149 | 102 | 43 | 10 |
|  | THORNY SKATE | SUMMER | 952 | 4,741 | 4,719 | 4,687 | 4,642 | 4,576 | 4,483 | 4,330 | 4,095 | 3,821 | 3,498 | 3,089 | 2,636 | 2,053 | 1,528 | 847 | 321 |
|  | THORNY SKATE | 1963-1971 | 527 | 2,627 | 2,617 | 2,607 | 2,587 | 2,554 | 2,504 | 2,437 | 2,329 | 2,199 | 2,050 | 1,862 | 1,627 | 1,324 | 1,086 | 660 | 275 |
|  | THORNY SKATE | 1972-1981 | 315 | 1,570 | 1,566 | 1,553 | 1,539 | 1,515 | 1,493 | 1,440 | 1,354 | 1,255 | 1,119 | 934 | 772 | 562 | 339 | 152 | 45 |
|  | THORNY SKATE | 1982-1991 | 35 | 174 | 171 | 169 | 168 | 165 | 160 | 157 | 150 | 146 | 134 | 116 | 91 | 64 | 31 | 9 | 0 |
|  | THORNY SKATE | 1992-2001 | 75 | 369 | 364 | 359 | 349 | 342 | 325 | 296 | 262 | 221 | 195 | 177 | 147 | 104 | 72 | 26 | 0 |
|  | THORNY SKATE | FAL | 3,659 | 18,194 | 18,090 | 17,923 | 17,687 | 17,342 | 16,831 | 16,030 | 14,937 | 13,700 | 12,420 | 10,676 | 9,031 | 6,884 | 4,928 | 2,952 | 1,212 |
|  | THORNY SKATE | 1963-1971 | 1,141 | 5,679 | 5,651 | 5,609 | 5,559 | 5,484 | 5,392 | 5,245 | 5,032 | 4,760 | 4,461 | 4,037 | 3,575 | 2,969 | 2,339 | 1,565 | 691 |
|  | THORNY SKATE | 1972-1981 | 1,627 | 8,103 | 8,067 | 8,005 | 7,913 | 7,769 | 7,553 | 7,162 | 6,642 | 6,008 | 5,388 | 4,509 | 3,696 | 2,675 | 1,790 | 947 | 347 |
|  | THORNY SKATE | 1982-1991 | 489 | 2,427 | 2,408 | 2,379 | 2,329 | 2,268 | 2,172. |  | 1,866 | 1,695 | 1,482 | 1,244 | 1,023 | 745 | 535 | 326 | 160 |
|  | THORNY SKATE | 1992-2001 | 284 | 1,408 | 1,396 | 1,377 | 1,351 | 1.100x | qex | Of1,69/ | 1,014 | 897 | 786 | 618 | 513 | 349 | 184 | 96 | 10 |
|  | THORNY SKATE | 2002-2012 | 118 | 576 | 567 | 554 | 536 | 517 | 484 | 434 | 383 | 339 | 303 | 269 | 224 | 146 | 80 | 19 | 3 |
|  | THORNY SKATE | All | 7,471 | 37,138 | 36,898 | 36,538 | 36,010 | 35,271 | 34,152 | 32,469 | 30,252 | 27,771 | 25,138 | 21,845 | 18,493 | 14,188 | 10,081 | 5,958 | 2,590 |

Synopsis of juvenile groundfish habitat and spawning analysis

| Approximate 20\% of biomass <br> (upper), 180 for maturity (lower) | Species | I Row Labels |  |  |  |  | $5^{5^{0^{0^{0}}}}=$ |  |  |  |  |  |  | $5^{55^{40}}$ | $5^{0^{40}}$ |  |  |  | $55^{5}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 75 cm | White hake | winter | 302 | 1,502 | 1,483 | 1,427 | 1,349 | 1,248 | 1,134 | 1,051 | 955 | 813 | 639 | 515 | 445 | 397 | 352 | 313 | 295 |
| L80 $=45 \mathrm{~cm}$ | White hake | 1963-1971 | 258 | 1,286 | 1,270 | 1,247 | 1,194 | 1,107 | 1,024 | 952 | 878 | 755 | 609 | 491 | 421 | 378 | 339 | 300 | 282 |
|  | White hake | 1972-1981 | 18 | 90 | 90 | 79 | 71 | 69 | 54 | 49 | 40 | 31 | 16 | 13 | 13 | 13 | 13 | 13 | 13 |
|  | White hake | 1992-2001 | 19 | 93 | 90 | 74 | 61 | 53 | 43 | 38 | 28 | 21 | 14 | 11 | 11 | 6 | 0 | 0 | 0 |
|  | White hake | 2002-2012 | 7 | 33 | 33 | 27 | 23 | 20 | 14 | 11 | 8 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | White hake | SPRING | 3,694 | 18,429 | 18,187 | 17,524 | 16,803 | 15,598 | 14,114 | 12,786 | 11,344 | 9,412 | 7,425 | 5,441 | 3,983 | 2,905 | 2,405 | 1,950 | 1,581 |
|  | WHITE HAKE | 1963-1971 | 170 | 849 | 839 | 816 | 769 | 690 | 614 | 561 | 506 | 432 | 364 | 321 | 273 | 240 | 212 | 171 | 138 |
|  | White hake | 1972-1981 | 1,691 | 8,445 | 8,358 | 8,125 | 7,843 | 7,410 | 6,813 | 6,296 | 5,769 | 5,008 | 4,198 | 3,157 | 2,331 | 1,610 | 1,320 | 1,118 | 961 |
|  | White hake | 1982-1991 | 795 | 3,967 | 3,900 | 3,712 | 3,538 | 3,270 | 2,966 | 2,698 | 2,346 | 1,919 | 1,413 | 981 | 695 | 572 | 494 | 422 | 356 |
|  | WHITE HAKE | 1992-2001 | 450 | 2,246 | 2,211 | 2,115 | 2,014 | 1,802 | 1,523 | 1,289 | 1,088 | 786 | 523 | 339 | 210 | 148 | 121 | 84 | 44 |
|  | White hake | 2002-2012 | 587 | 2,923 | 2,879 | 2,756 | 2,639 | 2,425 | 2,198 | 1,942 | 1,636 | 1,267 | 927 | 643 | 475 | 334 | 259 | 155 | 82 |
|  | White hake | SUMMER | 1,171 | 5,840 | 5,741 | 5,426 | 4,997 | 4,494 | 3,956 | 3,489 | 3,087 | 2,507 | 1,885 | 1,381 | 1,013 | 719 | 587 | 504 | 437 |
|  | White hake | 1963-1971 | 355 | 1,776 | 1,770 | 1,745 | 1,700 | 1,614 | 1,515 | 1,417 | 1,300 | 1,088 | 822 | 566 | 426 | 333 | 272 | 236 | 204 |
|  | WHITE HAKE | 1972-1981 | 414 | 2,070 | 2,062 | 1,998 | 1,861 | 1,722 | 1,561 | 1,416 | 1,290 | 1,089 | 884 | 715 | 537 | 369 | 316 | 268 | 233 |
|  | White hake | 1982-1991 | 135 | 672 | 652 | 562 | 436 | 343 | 247 | 174 | 124 | 73 | 32 | 20 | 9 | 0 | 0 | 0 | 0 |
|  | WHITE HAKE | 1992-2001 | 266 | 1,322 | 1,257 | 1,121 | 1,000 | 815 | 633 | 482 | 374 | 258 | 147 | 80 | 40 | 16 | 0 | 0 | 0 |
|  | WHITE HAKE | FAL | 5,519 | 27,377 | 26,873 | 26,313 | 24,673 | 22,062 | 19,488 | 17,049 | 14,531 | 11,918 | 9,129 | 6,826 | 5,143 | 3,764 | 2,940 | 2,370 | 1,933 |
|  | White hake | 1963-1971 | 779 | 3,885 | 3,826 | 3,725 | 3,542 | 3,217 | 2,909 | 2,616 | 2,284 | 1,899 | 1,509 | 1,136 | 897 | 716 | 651 | 528 | 490 |
|  | WHITE HAKE | 1972-1981 | 2,231 | 11,109 | 10,951 | 10,783 | 10,258 | 9,366 | 8,471 | 7,547 | 6,702 | 5,769 | 4,647 | 3,640 | 2,803 | 2,033 | 1,654 | 1,371 | 1,151 |
|  | WHITE HAKE | 1982-1991 | 1,080 | 5,307 | 5,164 | 5,020 | 4,548 | 3,881 | 3,308 | 2,822 | 2,313 | 1,840 | 1,354 | 960 | 628 | 402 | 243 | 182 | 142 |
|  | White hake | 1992-2001 | 801 | 3,968 | 3,891 | 3,798 | 3,537 | 3,120 | 2,646 | 2,188 | 1,705 | 1,237 | 788 | 533 | 412 | 329 | 231 | 168 | 73 |
|  | WHITE HAKE | 2002-2012 | 628 | 3,108 | 3,042 | 2,988 | 2,787 | 2,478 | 2,154 | 1,876 | 1,527 | 1,173 | 830 | 558 | 404 | 284 | 162 | 120 | 76 |
|  | WHITE HAKE | All | 10,687 | 53,149 | 52,284 | 50,691 | 47,823 | 43,402 | 38,693 | 34,375 | 29,917 | 24,650 | 19,078 | 14,164 | 10,583 | 7,784 | 6,285 | 5,138 | 4,247 |
| 30 cm | WINDOWPANE | WINTER | 1,033 | 4,331 | 1,304 | 119 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 = $\mathbf{2 5} \mathbf{~ c m}$ | WINDOWPANE | 1963-1971 | 28 | 134 | 77 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1972-1981 | 15 | 66 | 44 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1992-2001 | 869 | 3,573 | 978 | 79 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 2002-2012 | 121 | 557 | 205 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | SPRING | 834 | 3,681 | 1,863 | 426 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1963-1971 | 20 | 91 | 51 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1972-1981 | 439 | 1,948 | 948 | 186 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1982-1991 | 238 | 1,074 | 638 | 211 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1992-2001 | 75 | 306 | 124 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 2002-2012 | 62 | 262 | 102 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | SUMMER | 101 | 484 | 327 | 76 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1963-1971 | 19 | 94 | 67 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1972-1981 | 81 | 387 | 260 | 69 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1982-1991 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1992-2001 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | Fall | 1,097 | 4,636 | 2,200 | 420 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1963-1971 | 54 | 230 | 109 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1972-1981 | 370 | 1,668 | 955 | 200 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1982-1991 | 251 | 1,055 | 607 | 157 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 1992-2001 | 263 | 1,077 | 374 | 35 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | 2002-2012 | 159 | 607 | 155 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINDOWPANE | All | 3,066 | 13,132 | 5,695 | 1,041 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 cm | WINTER FLOUNDER | WINTER | 271 | 1,340 | 1,287 | 1,140 | 910 | 620 | 316 | 126 | 15 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L80 $=30 \mathrm{~cm}$ | WINTER FLOUNDER | 1963-1971 | 157 | 782 | 767 | 718 | 600 | 415 | 192 | 78 | 12 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1972-1981 | 43 | 214 | 209 | 188 | 165 | 132 | 87 | 40 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1992-2001 | 57 | 278 | 250 | 183 | 115 | 55 | 27 | 9 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 2002-2012 | 14 | 67 | 61 | 50 | 31 | 17 | 10 | O | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | SPRING | 2,113 | 9,986 | 8,765 | 6,791 | 4,642 | 2,690 | 1,090 | 344 | 94 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1963-1971 | 149 | 739 | 722 | 686 | 551 | 382 | 202 | 52 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1972-1981 | 650 | 3,164 | 2,906 | 2,392 | 1,698 | 1,003 | 431 | 169 | 53 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1982-1991 | 551 | 2,606 | 2,312 | 1,788 | 1,193 | 626 | 220 | 65 | 21 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1992-2001 | 279 | 1,323 | 1,161 | 834 | 535 | 271 | 96 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 2002-2012 | 484 | 2,154 | 1,663 | 1,092 | 665 | 408 | 141 | 34 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | SUMMER | 799 | 3,690 | 3,069 | 2,101 | 1,314 | 693 | 349 | 154 | 38 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1963-1971 | 159 | 794 | 776 | 709 | 564 | 305 | 140 | 62 | 18 | 3 | , | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1972-1981 | 529 | 2,437 | 1,978 | 1,274 | 709 | 382 | 208 | 92 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1982-1991 | 6 | 25 | 16 | 8 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1992-2001 | 105 | 434 | 300 | 110 | 39 | 6 | 0 | , | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | FAL | 3,111 | 14,859 | 12,977 | 9,244 | 5,730 | 3,254 | 1,584 | 584 | 153 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1963-1971 | 234 | 1,165 | 1,136 | 1,064 | 895 | 611 | 348 | 169 | 66 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1972-1981 | 762 | 3,719 | 3,392 | 2,690 | 1,858 | 1,095 | 575 | 225 | 52 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WIITER FLOUNDER | 1982-1991 | 396 | 1,857 | 1,579 | 1,097 | 664 |  | ${ }^{\frac{128}{128}}$ |  | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 1992-2001 | 812 | 3,868 | 3,282 | 1,969 | 997 | Ha | (20 ${ }^{\text {a }}$ | Of 69\% | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | 2002-2012 | 906 | 4,250 | 3,587 | 2,424 | 1,315 | 741 | 328 | 97 | 13 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | WINTER FLOUNDER | All | 6,294 | 29,876 | 26,098 | 19,277 | 12,596 | 7,257 | 3,339 | 1,208 | 301 | 53 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Synopsis of juvenile groundfish habitat and spawning analysis


Table 9. Summary of cluster analysis procedures applied to survey catch of juveniles (number) and large spawners (weight).

| Procedures run individually on age <br> $0 / 1$ juveniles ${ }^{1}$ and large spawners ${ }^{2}$ | Process | Sample size or effect |  |
| :--- | :--- | :--- | :--- |
| Hurdle model approach adjustment | Adjust cumulative catch at size, <br> multiplying by the proportion of | All tows included |  |
| Log transform | Transform non-zero catches to a <br> normalized distribution | Zero catches are ignored (reduced <br> number of tows analyzed) |  |
| Select tows for analysis | Select by survey, season, and <br> decade | Reduces number of tows; analysis <br> occurs in desired time period and <br> season; surveys analyzed <br> separately due to catchability <br> differences. Remaining tows may <br> be insufficient number to analyze <br> spatial autocorrelation or hotspots. |  |
| Spatial autocorrelation (Moran's I) | Determine range of highest spatial <br> autocorrelation to set Zone of <br> Indifference parameter for hotspot <br> analysis | Analyzes untransformed tows, <br> including zero catch tows. <br> Procedure may not detect a <br> significant positive spatial <br> autocorrelation. If peak is weak or <br> undetected by analysis, a <br> reasonable alternative was applied <br> for hot spot analysis. |  |
| Hot spot analysis (Getis-Ord's G*) <br> and selection | Identifies hotspots, filtered for <br> significant (p<0.05) hotspots above <br> the mean. | Procedure may not identify any <br> significant hotspots at p<0.05 level. |  |
| Grid hotspots | Number of significant hotspots for <br> a species within a 100 km ${ }^{2}$ SASI <br> grid is summed. | All surveys in a season are <br> included, since the hotspot data are <br> standardized relative to each <br> survey's mean. |  |
| Weight layers by importance factor | Number of hotspots in a grid is <br> multiplied by importance factor | Final grid for a season includes all <br> surveys where significant hotspots |  |

[^1]Synopsis of juvenile groundfish habitat and spawning analysis

| Procedures run individually on age <br> $0 / 1$ juveniles ${ }^{1}$ and large spawners ${ }^{2}$ | Process | Sample size or effect |  |
| :--- | :--- | :--- | :--- |
|  | and summed over species. | were identified by the analysis, <br> weighted by the relative <br> importance of the effect that spatial <br> management will have on regulated <br> groundfish. |  |

Table 10. Summary of peak spatial autocorrelation results and alternative trial peaks in parantheses. $\mathrm{NA}=$ analysis not attempted due to infrequent catch or data not yet available. $\mathrm{NP}=$ No significant peak autocorrelation detected. NSHS = No significant hotspots of above average catches detected or produced by the hotspot analysis. IC $=$ insufficient catch to conduct either a spatial autocorrelation or hotspot analysis.

|  | Survey: <br> NMFS spring |  | Survey: MADMF spring |  | Survey: ME/NH spring |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | Juvenile | Spawner | Juvenile | Spawner | Juvenile | Spawner |
| Cod | 8510 (11510) | 11510 | 10528 (15528) | 10525 (17528) | 4620 (10620) | 30620 |
| Haddock | 8010 (10010) | 8010 (20010) | 16528 | 10528 | 4620 (6620) | 13620 (NSHS) |
| Yellowtail flounder | 11510 | 11510 (16510) | 9528 (14528) | 8528 (17528) | IC | IC |
| American plaice | 14510 | 10510 | 8528 (17528) | 11528 | 15620 | 17620 |
| Atlantic wolffish | IC ( $2+$ tows) | NP (20010) | NA | NA | NA | NA |
| Ocean pout | 21510 (12 + tows) | 10510 | 15528 (22528) | 13528 | 5620 | 17620 NSHS |
| Pollock | 13510 | 10510 | NP (21 + tows) | IC | 3620 (7620) | IC |
| Red Hake | 11510 (14510) | NP (14510) | 8528 | 8528 | 9620 | 5620 |
| Redfish | 9510 | 10510 | IC | 11528 (NSHS) | 3620 (9620) | $\begin{gathered} 4620(17620) \\ \text { NSHS } \end{gathered}$ |
| Silver hake | 10510 | 32510 | 20639 | 10528 | 6620 | 11620 |
| White hake | NP (20010) | 8510 (21510) | NP (7528) | IC | 8620 | NP (10620) |
| Winter flounder | 11510 | 8510 (15510) | 7528 | 8528 | 3620 (14620) | $\begin{gathered} \text { NP 912620) } \\ \text { NSHS } \\ \hline \end{gathered}$ |
| Witch flounder | 13510 | 8510 | NP (8528) | IC | 7620 | NP (3620) NSHS |
| Windowpane flounder | 10510 (23510) | 8510 | 8528 NSHS | 8528 | 4320 NSHS | NP NSHS |
| Alewife | NA | NA | NA | NA | 7620 | 3620 (20620) |
| Atlantic herring | NA | NA | NA | NA | 4620 (7620) | 5620 (23620) |
| Atlantic halibut | NA | NA | NA | NA | 12620 | NP NSHS |
| Goosefish | NA | NA | NA | NA | NA | NA |
| Barndoor skate | NA | NA | NA | NA | NA | NA |


|  | Survey: IBS Cod spring |  | Survey: <br> IBS Goosefish spring |  | Survey: <br> NMFS dredge summer |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | Juvenile | Spawner | Juvenile | Spawner | Juvenile | Spawner |
| Cod | 4534 (13534) | NP (28534) | IC | 36226 | 10338 IC | IC |
| Haddock | 11534 | 7534 | NP (48226) NSHS | 34226 | 7338 (16338) | 9338 (13338) |
| Yellowtail flounder | IC | 13534 NSHS | IC | 34226 | 5338 | 5338 |
| American plaice | 6534 (9534) | 8534 | NA | NA | NA | NA |
| Atlantic wolffish | IC | IC | NA | NA | NA | NA |
| Ocean pout | IC | IC | NA | NA | NA | NA |
| Pollock | 5334 | 5334 IC | NA | NA | NA | NA |
| Red Hake | IC | IC | NA | NA | NP (19338) | IC |
| Redfish | 26534 (5534) | 2634 (5534) | NA | NA | NA | NA |
| Silver hake | IC | IC | NA | NA | NA | NA |
| White hake | 6534 (14534) | 6534 (14534) | NA | NA | NA | NA |
| Winter flounder | 5534 | 5534 | NA | NA | 16338 | 17338 |
| Witch flounder | 6534 NSHS | 6534 NSHS | NA | NA | NA | NA |
| Windowpane flounder | IC | IC | NA | NA | NA | NA |
| Alewife | NA | NA | NA | NA | NA | NA |
| Atlantic herring | NA | NA | NA | NA | NA | NA |
| Atlantic halibut | NA | NA | NA | NA | NA | NA |
| Goosefish | NA | NA | 35226 | NP | NP (19764) | 5338 (23338) |
| Barndoor skate | NA | NA | NA | NA | NP (15338) | 11338 (15338) |


|  | $\begin{array}{c}\text { Survey: }\end{array}$ |  | $\begin{array}{c}\text { Survey: } \\ \text { NMFS fall }\end{array}$ |  | $\begin{array}{c}\text { Survey: } \\ \text { MADMF fall }\end{array}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | Juvenile shrimp summer | Spawner | Juvenile | Spawner | Juvenile | Spawner |
| Cod | $8528(16528)$ | $7528(13528)$ | $8624(18624)$ | $8624(17624)$ | $7365(9365)$ | NP (5365) NSHS |
| Haddock | 8528 | $20528(26528)$ | 13624 | 13624 | 6365 (strong SAC) | 22365 |
| Yellowtail flounder | NA | NA | 9624 | 14264 | NP (31365) NSHS | $4365(22365)$ |
| NASHS |  |  |  |  |  |  |$]$


|  | Survey: ME/NH fall |  | Survey: IBS Cod fall |  | Survey: IBS YTF fall |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | Juvenile | Spawner | Juvenile | Spawner | Juvenile | Spawner |
| Cod | 5988 (7988) | 4988 (21998) | 7313 | 9313 | IC | IC |
| Haddock | 29998 | NP IC | 7313 | 20913 | IC | IC |
| Yellowtail flounder | 8988 NSHS | NP IC | IC | 5313 | 24642 NSHS | 16642 |
| American plaice | 24988 | 3988 | 5313 | NP (25313) | NA | NA |
| Atlantic wolffish | NA | NA | IC | IC | NA | NA |
| Ocean pout | 4998 | IC | NA | NA | NA | NA |
| Pollock | NP (18998) | IC | NP (11313) NSHS | 12313 | NA | NA |
| Red Hake | 16998 (strong peak) | 10998 (strong peak) | IC | IC | NA | NA |
| Redfish | 5998 (17998) | NP 6998 | 12313 | NP (8313) | NA | NA |
| Silver hake | 13998 | 9988 | IC | IC | NA | NA |
| White hake | 17998 | 6998 IC | 10313 | IC | NA | NA |
| Winter flounder | 17998 | NP IC | 5313 (17313) | 7313 | IC | IC |
| Witch flounder | 4998 (14998) | $\begin{gathered} 8998(17998) \\ \text { NSHS } \end{gathered}$ | NP | 5313 (9313) | NA | NA |
| Windowpane flounder | 8988 | 3988 IC | IC | 7313 | NA | NA |
| Alewife | 16988 | 7988 (17988) | NA | NA | NA | NA |
| Atlantic herring | 5998 | 3988 | NA | NA | NA | NA |
| Atlantic halibut | 12998 IC | 3998 IC | NA | NA | NA | NA |
| Goosefish | 11998 NSHS | IC | 5313 (9313) | NP (23313) | NP IC | IC |
| Barndoor skate | NA | NA | NA | NA | NA | NA |


|  | Survey: NMFS winter |  | Survey: IBS Cod winter |  | Survey: IBS GSF winter |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | Juvenile | Spawner | Juvenile | Spawner | Juvenile | Spawner |
| Cod | 15806 | 27806 | 9728 (12728) | NP (7728) | NP (31083) NSHS | NP |
| Haddock | 17806 | NP (23806) | 17728 (31728) | 10728 | NP | 49083 |
| Yellowtail flounder | 21806 | 12806 (28806) | IC | NP (3728) | IC | NP |
| American plaice | IC | 24806 | 8728 | 6728 | 59083 NSHS | 35083 NSHS |
| Atlantic wolffish | NA | NA | IC | IC | NA | NA |
| Ocean pout | 14806 (16806) | 14806 | IC | IC | NA | NA |
| Pollock | IC | IC | IC | NP (15728) | NA | NA |
| Red Hake | 20806 (27806) | 12806 | NA | NA | NA | NA |
| Redfish | NA | NA | NA | NA | NA | NA |
| Silver hake | 19806 | 12806 (31806) | NA | NA | NA | NA |
| White hake | NA | NA | 11728 | NP IC |  |  |
| Winter flounder | 12806 (16806) | 21806 | 5728 (20728) | NP (24728) NSHS | 35083 | NP NSHS |
| Witch flounder | 19806 | 12806 (14806) | 7728 (12728) | 8728 | IC | 36083 (40083) |
| Windowpane flounder | 15806 (17806) | 14806 (37806) | IC | 6728 | NA | NA |
| Alewife | NA | NA | NA | NA | NA | NA |
| Atlantic herring | NA | NA | NA | NA | NA | NA |
| Atlantic halibut | NA | NA | NA | NA | NA | NA |
| Goosefish | 12806 (25806) | 32806 | 6728 (21728) | NP | 35083 (44083) | 34083 |
| Barndoor skate | NA | NA | NA | NA | 40083 NSHS | NP NSHS |

## Synopsis of juvenile groundfish habitat and spawning analysis

Table 11. Summary of significant hotspots of above average catches identified by survey and species for age $0 / 1$ juvenile (upper) and for large spawners (lower), $2002-2012$.

| Survey | Years | Tows | Mean to ne: | StdDev | 90th pctle | 95th pctle | 4/evi/s |  |  |  |  |  |  | ose/s/n | ${ }^{1 / 20} \mathrm{OH}_{\mathrm{O}}$ |  |  |  | $P_{0}{ }_{0}$ |  | B/2/or |  |  | e/ome |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NMFS spring | 2002-2012 | 3,426 | 4,012.0 | 3,630.0 | 7,509.5 | 9,014.9 |  | 85 | 0 |  |  |  | 35 |  | 31 | 0 | 0 | 122 | 25 | 167 | 70 | 53 | 7 | 3 | 11 | 609 |
| NMFS shrimp |  | 677 | 3,088.9 | 2,328.5 | 6,527.5 | 8,258.9 |  | 114 |  |  |  |  | 1 | 48 | 4 |  |  | 23 | 161 | 87 | 112 |  | 56 |  |  | 606 |
| NMFS scallop | 2002-2011 | 4,634 | 1,538.7 | 1,454.9 | 3,337.7 | 4,269.8 |  |  |  |  |  | 81 | 18 | 250 | 61 |  |  | 0 |  |  |  | 14 |  |  | 7 " | 431 |
| NMFSfall | 2002-2011 | 3,413 | 4,004.0 | 2,634.0 | 7,624.0 | 8,979.0 |  | 91 | 1 |  |  |  | 33 | 30 | 80 | 0 | 1 | 286 | 69 | 254 | 77 | 132 | 19 | 4 | 5 | 082 |
| NMFS winter | 2002-2007 | 659 | 6,212.4 | 5,272.9 | 11,805.6 | 13,468.3 |  | 0 |  |  |  |  | 2 | 3 | 1 | 1 |  | 18 |  | 59 |  | 8 | 3 | 4 | 0 | 99 |
| MADMF spring | 2002-2012 | 936 | 832.9 | 655.3 | 1,798.9 | 2,184.9 |  | 44 |  |  |  |  | 80 |  | 8 | 0 | 3 | 19 | 0 | 41 | 4 | 150 |  | 0 | $17^{\prime \prime}$ | 366 |
| MADMF fall | 2002-2011 | 714 | 1,096.8 | 835.9 | 2,364.8 | 2,807.9 |  | 24 | 1 |  |  |  | 5 | 0 | 4 | 0 | 0 | 58 | 0 | 88 | 2 | 131 |  | 2 |  | 315 |
| MENH spring |  | 1,194 | 1,078.7 | 1,156.7 | 2,619.4 | 3,298.2 | 187 | 269 | 51 | 19 |  |  | 85 |  | 36 | 9 | 16 | 70 | 116 | 317 | 71 | 264 | 57 | 149 | 0 | 716 |
| MENHfall |  | 812 | 1,271.7 | 1,436.0 | 2,987.9 | 3,859.1 | 192 | 233 | 92 | 11 |  |  | 29 | 0 | 15 | 4 | 4 | 186 | 329 | 275 | 209 | 187 | 46 | 134 | 0 | 946 |
| IBS cod spring |  | 449 | 1,513.1 | 1,643.0 | 3,533.9 | 4,638.3 |  | 77 |  |  |  |  | 54 |  | 25 |  |  |  | 18 |  | 10 | 16 | 0 |  |  | 200 |
| IBS cod fall |  | 175 | 2,202.4 | 2,559.9 | 4,312.8 | 6,101.3 |  | 12 |  |  |  |  | 21 | 7 | 8 |  | 0 |  | 2 |  | 8 | 28 | 0 |  |  | 86 |
| IBS cod winter |  | 274 | 2,064.9 | 3,114.4 | 3,728.0 | 5,131.3 |  |  |  |  |  |  | 2 | 10 | 10 |  |  |  |  |  | 14 | 65 | 1 |  |  | 102 |
| IBS goosefish spring |  | 229 | 15,551.0 | 13,125.6 | 30,226.1 | 34,028.5 |  |  |  |  |  |  |  | 13 | 1 |  |  |  |  |  |  |  |  |  |  | 13 |
|  |  | 198 | 16,992.9 | 9,778.9 | 31,082.6 | 34,286.3 |  |  |  |  |  |  | 2 |  | 0 |  |  |  |  |  |  |  |  |  |  |  |
| IBSYTF fall |  | 709 | 3,382.5 | 14,471.1 | 5,642.0 | 7,373.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $0{ }^{*}$ | 0 |
|  |  |  |  |  | Total species hotspots $=$ |  | 379 | 949 | 145 | 30 | 0 | 81 | 367 | 361 | 283 | 14 | 24 | 782 | 720 | 1288 | 577 | 1048 | 189 | 296 | 40 | 7573 |



Figure 8. Data processing flowchart for spatial autocorrelation and hotspot analyses for juvenile (upper) and large spawner (lower) life stages. The example analyzes witch flounder juvenile and large spawner distribution in the 2009 IBS winter goosefish survey.


Figure 9. Workflow for merging and gridding weighted number of hotspots for a season.

Synopsis of juvenile groundfish habitat and spawning analysis


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 10. Juvenile cod (<= 25 cm ) per tow in 2002-2012 NMFS spring trawl surveys vs. Getis-Ords G* hotspot statistics for $\mathbf{2 2 9}$ hotspots derived from 3426 tow locations. All tows are non-zero and the diameter is scaled to untransformed catch per tow. Low $p$ values represent significant clusters. Positive $Z$ scores are above the mean of non-zero tows. Tows that fall within the light blue box represent high catch rates derived from significant ( $\mathrm{p}<=0.05$ ) clusters.


Map 1. Location of above average significant hotspots (blue circles) compared to all clusters (shaded circles) overlaying scaled $<=25 \mathrm{~cm}$ cod/tow (pink squares), NMFS spring trawl survey 2002-2012.


Synopsis of juvenile groundfish habitat and spawning analysis
Figure 11. Presence (red)/absence (red) of cod in spawning condition observed during the 2002-2012 NMFS spring trawl surveys.


Synopsis of juvenile groundfish habitat and spawning analysis
Figure 12. Presence (red)/absence (red) of haddock in spawning condition observed during the 2002-2012 NMFS spring trawl surveys.


Synopsis of juvenile groundfish habitat and spawning analysis
Figure 13. Presence (red)/absence (red) of haddock in spawning condition observed during the 2002-2012 NMFS spring trawl surveys.


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 14. Coastal juvenile groundfish habitat management area option, compared to a summary grid of weighted hotspots (darker shade denotes a higher weighted hotspot value; outlined and unshaded blocks represent areas with hotspots given zero weight).


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 15. Juvenile groundfish habitat management area option, compared to a summary grid of weighted hotspots (darker shade denotes a higher weighted hotspot value; outlined and unshaded blocks represent areas with hotspots given zero weight).


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 16. Seasonal groundfish spawning areas derived from hotspot analysis.


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 17. Proposed March-April modified rolling closure option (black outline) compared to existing April sector rolling closure (shaded).


Figure 18. Proposed May modified rolling closure option (black outline) compared to existing May sector rolling closure (shaded).


Figure 19. Proposed June modified rolling closure option (black outline) compared to existing June sector rolling closure (shaded).


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## Getis-Ord Gi* statistic in ArcGIS

The Hot Spot Analysis tool calculates the Getis-Ord Gi* statistic (pronounced G-i-star) for each feature in a dataset. The resultant z-scores and p-values tell you where features with either high or low values cluster spatially. This tool works by looking at each feature within the context of neighboring features. A feature with a high value is interesting but may not be a statistically significant hot spot. To be a statistically significant hot spot, a feature will have a high value and be surrounded by other features with high values as well. The local sum for a feature and its neighbors is compared proportionally to the sum of all features; when the local sum is very different from the expected local sum, and that difference is too large to be the result of random chance, a statistically significant z-score results.

## Calculations

The Getis-Ord local statistic is given as:

$$
\begin{equation*}
G_{i}^{*}=\frac{\sum_{j=1}^{n} w_{i, j} x_{j}-\bar{X} \sum_{j=1}^{n} w_{i, j}}{S \sqrt{\frac{\left[n \sum_{j=1}^{n} w_{i, j}^{2}-\left(\sum_{j=1}^{n} w_{i, j}\right)^{2}\right]}{n-1}}} \tag{1}
\end{equation*}
$$

where $x_{j}$ is the attribute value for feature $j, w_{i, j}$ is the spatial weight between feature $i$ and $j, n$ is equal to the total number of features and:

$$
\begin{align*}
\bar{X} & =\frac{\sum_{j=1}^{n} x_{j}}{n}  \tag{2}\\
S & =\sqrt{\frac{\sum_{j=1}^{n} x_{j}^{2}}{n}-(\bar{X})^{2}}
\end{align*}
$$

The $G_{i}^{*}$ statistic is a $z$-score so no further calculations are required.

## Interpretation

The Gi* statistic returned for each feature in the dataset is a z-score. For statistically significant positive $z$-scores, the larger the $z$-score is, the more intense the clustering of high values (hot spot). For statistically significant negative $z$-scores, the smaller the $z$-score is, the more intense the clustering of low values (cold spot). For more information about determining statistical significance, see What is a $z$-score? What is a p-value?

## Output

This tool creates a new Output Feature Class with a z-score and p-value for each feature in the Input Feature Class. If there is a selection set applied to the Input Feature Class, only selected features will be analyzed, and only selected features will appear in the Output Feature Class. This tool also returns the $z$-score and p-value field names as derived output values for potential use in custom models and scripts.

## Synopsis of juvenile groundfish habitat and spawning analysis

When this tool runs in ArcMap, the Output Feature Class is automatically added to the table of contents with default rendering applied to the $z$-score field. The hot to cold rendering applied is defined by a layer file in <ArcGIS>/ArcToolbox/Templates/Layers. You can reapply the default rendering, if needed, by importing the template layer symbology.

## Hot spot analysis considerations

There are three things to consider when undertaking any hot spot analysis:

1. What is the Analysis Field (Input Field)? The hot spot analysis tool assesses whether high or low values (the number of crimes, accident severity, or dollars spent on sporting goods, for example) cluster spatially. The field containing those values is your Analysis Field. For point incident data, however, you may be more interested in assessing incident intensity than in analyzing the spatial clustering of any particular value associated with the incidents. In that case, you will need to aggregate your incident data prior to analysis. There are several ways to do this:

- If you have polygon features for your study area, you can use the Spatial Join tool to count the number of events in each polygon. The resultant field containing the number of events in each polygon becomes the Input Field for analysis.
- Use the Create Fishnet tool to construct a polygon grid over your point features. Then use the Spatial Join tool to count the number of events falling within each grid polygon. Remove any grid polygons that fall outside your study area. Also, in cases where many of the grid polygons within the study area contain zeros for the number of events, increase the polygon grid size, if appropriate, or remove those zero-count grid polygons prior to analysis.
- Alternatively, if you have a number of coincident points or points within a short distance of one another, you can use Integrate with the Collect Events tool to (1) snap features within a specified distance of each other together, then (2) create a new feature class containing a point at each unique location with an associated count attribute to indicate the number of events/snapped points. Use the resultant ICOUNT field as your Input Field for analysis.


## ENote:

If you are concerned that your coincident points may be redundant records, the Find Identical tool can help you to locate and remove duplicates.


Strategies for aggregating incident data
2. Which Conceptualization of Spatial Relationships is appropriate? What Distance Band or Threshold Distance value is best?
The recommended (and default) Conceptualization of Spatial Relationships for the Hot Spot Analysis (Getis-Ord Gi*) tool is Fixed Distance Band. Space-Time Window, Zone of Indifference, Contiguity, K Nearest Neighbor, and Delaunay Triangulation may also work well. For a discussion of best practices and strategies for determining an analysis distance value, see Selecting a Conceptualization of Spatial Relationships and Selecting a Fixed Distance. For more information about space-time hot spot analysis, see Space-Time Analysis.
3. What is the question?

This may seem obvious, but how you construct the Input Field for analysis determines the types of questions you can ask. Are you most interested in determining where you have lots of incidents, or where high/low values for a particular attribute cluster spatially? If so, run Hot Spot Analysis on the raw values or raw incident counts. This type of analysis is particularly helpful for resource allocation types of problems. Alternatively (or in addition), you may be interested in locating areas with unexpectedly high values in relation to some other variable. If you are analyzing foreclosures, for example, you probably expect more foreclosures in locations with more homes (said another way, at some level, you expect the number of foreclosures to be a function of the number of houses). If you divide the number of foreclosures by the number of homes, then run the Hot Spot Analysis tool on this ratio, you are no longer asking Where are there lots of foreclosures?; instead, you are asking Where are there unexpectedly high numbers of foreclosures, given the number of homes? By creating a rate or ratio prior to analysis, you can control for certain expected relationships (for example, the number of crimes is a function of population; the number of foreclosures is a function of housing stock) and identify unexpected hot/cold spots.

## Best practice guidelines

- Does the Input Feature Class contain at least 30 features? Results aren't reliable with less than 30 features.
- Is the Conceptualization of Spatial Relationships you selected appropriate? For this tool, the Fixed Distance Band method is recommended. For space-time hot spot analysis, see Selecting a Conceptualization of Spatial Relationships.
- Is the Distance Band or Threshold Distance appropriate? See Selecting a Fixed Distance.

B All features should have at least one neighbor.
B No feature should have all other features as neighbors.
B Especially if the values for the Input Field are skewed, you want features to have about eight neighbors each.

## Potential applications

Applications can be found in crime analysis, epidemiology, voting pattern analysis, economic geography, retail analysis, traffic incident analysis, and demographics. Some examples include the following:

- Where is the disease outbreak concentrated?
- Where are kitchen fires a larger than expected proportion of all residential fires?
- Where should the evacuation sites be located?
- Where/When do peak intensities occur?
- Which locations and at during what time periods should we allocate more of our resources?


## Additional resources

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## How Incremental Spatial Autocorrelation works in ArcGIS


#### Abstract

Desktop » Geoprocessing » Tool reference » Spatial Statistics toolbox » Analyzing Patterns toolset With much of the spatial data analysis you do, the scale of your analysis will be important. The default Conceptualization of Spatial Relationships for the Hot Spot Analysis tool, for example, isFIXED_DISTANCE_BAND and requires you to specify a distance value. For many density tools you will be asked to provide a Radius. The distance you select should relate to the scale of the question you are trying to answer or to the scale of remediation you are considering. Suppose, for example, you want to understand childhood obesity. What is your scale of analysis? Is it at the individual household or neighborhood level? If so, the distance you use to define your scale of analysis will be small, encompassing the homes within a block or two of each other. Alternatively, what will be the scale of remediation? Perhaps your question involves where to increase after-school fitness programs as a way to potentially reduce childhood obesity. In that case, your distance will likely be reflective of school zones. Sometimes it's fairly easy to determine an appropriate scale of analysis; if you are analyzing commuting patterns and know that the average journey to work is 12 miles, for example, then 12 miles would be an appropriate distance to use for your analysis. Other times it is more difficult to justify any particular analysis distance. This is when the Incremental Spatial Autocorrelation tool is most helpful. Whenever you see spatial clustering in the landscape, you are seeing evidence of underlying spatial processes at work. Knowing something about the spatial scale at which those underlying processes operate can help you select an appropriate analysis distance. The Incremental Spatial Autocorrelation tool runs the Spatial Autocorrelation (Global Moran's I) tool for a series of increasing distances, measuring the intensity of spatial clustering for each distance. The intensity of clustering is determined by the z-score returned. Typically, as the distance increases, so does the z-score, indicating intensification of clustering. At some particular distance, however, the z-score generally peaks. Sometimes you will see multiple peaks.




Peaks reflect distances where the spatial processes promoting clustering are most pronounced. The color of each point on the graph corresponds to the statistical significance of the $\underline{z}$-score values.

| Significance Level <br> (p-value) |  | Critical Value <br> (z-score) |
| ---: | :--- | :--- |
| 0.01 | $\square$ | $<-2.58$ |
| 0.05 | $\square$ | $-2.58--1.96$ |
| 0.10 | $\square$ | $-1.96--1.65$ |
| - | $\square$ | $-1.65-1.65$ |
| 0.10 | $\square$ | $1.65-1.96$ |
| 0.05 | $\square$ | $1.96-2.58$ |
| 0.01 | $\square$ | $>2.58$ |

One strategy for identifying an appropriate scale of analysis is to select the distance associated with the statistically significant peak that best reflects the scale of your question. Often this is the first statistically significant peak.

## How do I select the Beginning Distance and Distance Increment values?

All distance measurements are based on feature centroids and the default Beginning Distance is the smallest distance that will ensure every feature has at least one neighboring feature. This is generally a good choice, unless your dataset includes spatial outliers. Determine whether or not you have spatial outliers, then select all but the outlier features and run Incremental Spatial Autocorrelation on just the selected features. If you find a peak distance for the selection set, use that distance to create a spatial weights matrix file based on all of your features (even the outliers). When you run theGenerate Spatial Weights Matrix tool to create the spatial weights matrix file, set the Number of

## Synopsis of juvenile groundfish habitat and spawning analysis

Neighbors parameter to some value so that all features will have at least that many neighboring features.
The default Increment Distance is the average distance to each feature's nearest neighboring feature. If you've determined an appropriate starting distance using the strategies above and still don't see a peak distance, you may want to experiment with smaller or larger increment distances.

## What if the graph never peaks?

In some cases, you will use the Incremental Spatial Autocorrelation tool and get a graph with a zscore that just continues to rise with increasing distances; there is no peak. This most often happens in cases where data has been aggregated and the scale of the processes impacting your input Field variable are smaller than the aggregation scheme. You can try making your Distance Increment smaller to see if this captures more subtle peaks. Sometimes, however, you won't get a peak because there are multiple spatial processes, each operating at a different distance, in your study area. This is often the case with large point datasets that are noisy (no clear spatial pattern to the point data values you're analyzing). In this case, you will need to justify your scale of analysis using some other criteria.

## Interpreting results

When you run the Incremental Spatial Autocorrelation tool in the foreground, the z-score results for each distance are written to the Progress window. This output is also available from the Results window. If you right-click on the Messages entry in the Results window and select View, the tool results are displayed in a Message dialog box. When you specify a path for the optional Output Table parameter, a table is created that includes fields
for Distance, Moransl, Expectedl, Variance, z_score, and p_value. By examining the z-score values in the Progress window, Message dialog box, or Output Table, you can determine if there are any peak distances. More typically, however, you would identify peak distances by looking at the graphic in the optional Output Report file. The report has three pages. An example of the first page of the report is shown below. Notice that this graph has three peak z-scores associated with distances of 5000,9000 , and 13000 feet. A halo will be drawn to highlight both the first peak distance and the maximum peak distance, but all peaks represent distances where the spatial processes promoting clustering are most pronounced. You can select the peak that best reflects the scale of your analytical question. In some cases, there will only be one halo because the first and the maximum peaks are found at the same distance. If none of the $z$-score peaks are statistically significant, then none of the peaks will have the light blue halo. Notice that the color of the plotted $z$-score corresponds to the legend showing the critical values for statistical significance.

## Spatial Autocorrelation by Distance



On page two of the report, the distances and $z$-score values are presented in table format. The last page of the report documents the parameter settings used when the tool was run. To get a report file, provide a path for the Output Report parameter.

Figure 20. Example of 'good' spatial autocorrelation result: Large spawner silver hake from MADMF fall survey, 20022011.

## Spatial Autocorrelation by Distance



Figure 21. Example of 'satisfactory' spatial autocorrelation result, with secondary peak autocorrelation: Juvenile American plaice from IBS cod fall survey, 2002-2011.

Spatial Autocorrelation by Distance


## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 22. Example of unsatisfactory spatial autocorrelation result, with no significant peak in autocorrelation: Large spawner American plaice from IBS cod fall survey, 2002-2011. In this case, hotspot analysis was re-run with a zone of indifference parameter of 25313 m , corresponding of a secondary non-significant spatial autocorrelation peak, but there were no significant hotspots identified nonetheless.

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Figure 23. Example of unsatisfactory spatial autocorrelation resulting from insufficient non-zero catches: Large spawner pollock from IBS cod fall survey, 2002-2011. No significant hotspots were identified and no further analysis was attempted.

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Figure 24. Example of 'good' spatial autocorrelation result, but first autocorrelation peak is probably not meaningful: Juvenile winter flounder from IBS cod fall survey, 2002-2011. The maximum peak of $17,313 \mathrm{~m}$ was used as the Zone of Indifference parameter in the hotspot analysis in lieu of the first peak.

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Figure 25. Example of unsatisfactory spatial autocorrelation: Juvenile witch flounder from IBS cod fall survey, 20022011. No significant hotspots were identified and no further analysis was attempted.

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## Synopsis of juvenile groundfish habitat and spawning analysis

Figure 26. Example of 'good' spatial autocorrelation result, with no meaningful first autocorrelation: Large spawner yellowtail flounder from NMFS winter survey, 2002-2007. The maximum peak was applied as a Zone of Indifference parameter in the hotspot analysis.

## Spatial Autocorrelation by Distance



Figure 27. Example of 'poor' spatial autocorrelation result. Data are sparse and tend the spatial autocorrelation has a 'choppy' appearance: Juvenile cod from NMFS winter survey, 2002-2007. Usually, this pattern is associated with a hotspot analysis that has no significant positive hotspots.

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Figure 28. Example of 'strong' spatial autocorrelation result: Large spawner witch flounder from the NMFS winter survey, 2002-2007.

Spatial Autocorrelation by Distance



[^0]:    ${ }^{1}$ Either SSBmsy/SSB or Bmsy/B used depending on what is reported in the assessment
    2Derived from Table 81 in Framework 48 or from NEFSC biological data. 1=no subpopulations, 2=some evidence, 3=known subpopulations
    3Based on information in literature. 1=less resident, more migratory; 2=more resident, less migratory
    4Sums include a mean value for unknowns

[^1]:    ${ }^{1}$ For aged species, upper size threshold that approximated $90^{\text {th }}$ percentile of age 1 fish. Threshold set at the approximate L20 for maturity for unaged species.
    ${ }^{2}$ Lower size threshold set where fish at or larger than the threshold comprised $20 \%$ of estimated biomass in the spring (applied to spring and summer) and fall (applied to fall and winter) NMFS trawl surveys.
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